Site-Specific Critical Loads of Acid Deposition on Soils in the Edmonton 83H East Map Sheet, Alberta

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EXECUTIVE SUMMARY

The Alberta Acid Deposition Management Framework stipulates that Alberta Environment is responsible for conducting an evaluation of the acid deposition monitoring data in the province, as well as conducting an evaluation of receptor sensitivity. The present study addresses the latter of these two requirements with respect to potentially acid sensitive soils in the Edmonton NTS 83H East Half map sheet. This map sheet extends from 112°W longitude to 113°W longitude, and from 53°N latitude to 54°N latitude.

The objective of this receptor sensitivity study is to provide an estimate of the critical load for the soil types and water bodies in the Edmonton East study area. On the basis of these critical loads estimates, a recommendation regarding the sensitivity of the study area as a whole is provided.

This area is herein referred to as the Edmonton East grid cell. The areal extent of the study area is approximately 7,386 square kilometres.

Three sensitivity assessment and modelling approaches were applied in examination of the soils of the Edmonton East grid cell. Critical load assessment by the empirical method referred to as the Skokloster approach resulted in a wide range of critical loads for soils ranging in texture from sands to clays. The method is not specifically applied to soil series, but to textural groupings of soils. This approach indicated critical loads as follows: very coarse textured soils - 0.2-0.5 kmol ha⁻¹ yr⁻¹; moderately coarse textured soils - 0.5-1.0 kmol ha⁻¹ yr⁻¹; medium to moderately fine textured soils - 1.0-2.0 kmol ha⁻¹ yr⁻¹; fine textured soils >2.0 kmol ha⁻¹ yr⁻¹.

The Steady State Mass Balance approach treats the soil as a single compartment to a 0.75 m depth. The SSMB assessment resulted in relatively high critical loads, as follows: very coarse textured soils - 0.6-0.7 kmol ha⁻¹ yr⁻¹; moderately coarse to medium textured soils – 1.0-1.6 kmol ha⁻¹ yr⁻¹; medium to moderately fine textured soils - 2.9 kmol ha⁻¹ yr⁻¹; and fine textured soils – 5.6 kmol ha⁻¹ yr⁻¹.

The ARC model utilizes the buffering capacity of soils due to cation exchange as well as to weathering, and assesses changes in soil chemistry over time. The modelling results were expressed as critical loads, which were subsequently applied in deriving sensitivity classes of soils. Two soil series (Primula and Nestow) were determined to be potentially (highly) sensitive to acid deposition in some places. Some of the samples of these series showed Moderate Sensitivity to acidification. Both these soils are Brunisols developed on very coarse (sandy) materials. Soils of coarse to moderately coarse texture (sand to loamy sand; Helliwell and Mundare soil series) showed Moderate Sensitivity in some soil samples, and Low sensitivity in others. The differences in sensitivity within the same soil series are thought to be related to the amount of organic matter in the A horizon. The north part of the Edmonton East grid cell is located in a transition area between Chernozemic soils to the south and forested Brunisolic and Luvisolic soils to the north. It is likely that those soils with relatively low amounts of organic matter are the most highly Sensitive to acid deposition. All other soils in the grid cell, being of

finer texture and having A horizons rich in organic matter, were rated based on the basis of previous studies as being of Low sensitivity.

Acidification sensitivity categories of soils examined in this study were compared to sensitivity classes in mapping carried out by Holowaychuk and Fessenden (1987). In the Holowaychuk-Fessenden mapping, only the Cooking Lake moraine (i.e., the Islet Upland Land System) was identified as having soils with potentially medium sensitivity. All other soils were categorized as being of Low Sensitivity to acidification. The ARC modelling results suggest that the predominantly Luvisolic soils of the Islet Upland have Low sensitivity to acidification. Other differences between the Holowaychuk-Fessenden mapping and the ARC model results pertain to the sandy Brunisolic soils of areas such as Redwater Plain, Eldorena Plain and Musidora Upland, in the north part of the Edmonton East grid cell. These are mapped as being of Low Sensitivity in the Holowaychuk-Fessenden map. In the current study, some of the soils that characterize these Land Systems were indicated as being Sensitive or of Moderate sensitivity according to the ARC model.

A map depicting the Land Systems, land cover and soil sensitivity to acid inputs in the Edmonton East map sheet was developed based on the soil sensitivity assessment and on land cover information. Proportions of soil series within Land Systems were estimated from information provided in AGRASID, and from this, the proportions of soils in Moderate to Sensitive (Nestow and Primula), Moderate to Low (Helliwell and Mundare), and Low (all other soils) acidification sensitivity categories were derived. The assignment of Sensitive, Moderate and Low Sensitivity categories was applied only to lands classified as having grassland, tree or shrub cover, on the basis of land use mapping by the Prairie Farm Rehabilitation Administration in 1993-1995. Cultivated soils were not rated, nor were lands categorized as 'Other Lands'. Three sensitivity map units were developed: Low Sensitivity, Low-Moderate Mix, and Low-Moderate-Sensitive Mix.

Portions of five land systems in the Edmonton 83H East Half grid cell were characterized as having a component of Sensitive and Moderately Sensitive soils. These are the Eldorena Plain, Redwater Plain, Musidora Upland, Edward Upland and Norma Plain. Other Land Systems likely have small components of Sensitive and Moderately Sensitive soils, but of too low extent to enable mapping at the scale applied in this assessment. Sensitive soils account for 0.5% and Moderately Sensitive soils account for 1.3% of the entire grid cell area. As defined in the Acid Deposition Management Framework (Clean Air Strategic Alliance and Alberta Environment 1999), this finding does not support the assignment of this grid cell to a Sensitive or Moderate Sensitivity rating.

1.0 INTRODUCTION

1.1 BACKGROUND

The Acid Deposition Management Framework for the long-term, provincial management of acid deposition (Clean Air Strategic Alliance and Alberta Environment 1999) was implemented in December, 1999. This framework is based upon the current understanding of the levels of acid deposition and the sensitivity of soil and water receptors in the province. Development of this framework included significant stakeholder consultation through Alberta's Clean Air Strategic Alliance.

Critical loads are the foundation of the framework. A critical load is a property of the receptor (soil, water), and is defined as the amount of acid input that can be received by the receptor that will not cause chemical changes leading to long-term harmful change to the receptor.

The province of Alberta is divided into grid cells measuring 1° latitude x 1° longitude, and each grid cell is categorized as being Sensitive, Moderately sensitive or of Low sensitivity to acid deposition based upon soil and water sensitivity databases. A Sensitive cell is defined as a cell within which 5% or more of the area is categorized as being Sensitive, and to such cells, a critical load of 0.25 kmol H⁺ ha⁻¹ yr⁻¹ is applied. A Moderately sensitive cell is defined as a cell within which less than 5% of the area is categorized as Sensitive, but where the total of Sensitive and Moderately sensitive areas equals or exceeds 5% of the cell area. To these Moderately sensitive cells, a critical load of 0.50 kmol H⁺ ha⁻¹ yr⁻¹ applied. The remainder of the grid cells are classified as being of Low sensitivity to acid deposition and are assigned a critical load of 1.00 kmol H⁺ ha⁻¹ yr⁻¹.

In addition to critical loads, grid cells have also been assigned target and monitoring loads. Target loads are based upon the critical loads, with the added proviso that target loads be an expression of society's values - in the Alberta framework, target loads are set at 90% of the critical loads (0.22, 0.45 and 0.90 kmol H⁺ ha⁻¹ yr⁻¹ for the three sensitivity classes). These target loads are also the environmental objectives as defined in provincial environmental legislation. By establishing target loads below the critical loads, provincial stakeholders and Alberta Environment have established a system of preventing an increase in deposition to the level believed harmful (the critical load). An exceedance of a target load will initiate processes to reduce emissions such that deposition in the exceedance cell is reduced to or below the target load for that cell.

Monitoring loads are also assigned to the sensitivity classes; these are set at 70% of the critical loads. Exceedance of this load initiates studies of receptor sensitivity and monitoring of deposition - the results of such studies are used to revise the initial assignments of cell sensitivity (and therefore the assigned numerical loads). If the studies confirm model prediction and sensitivity, the cell is watched more closely to ensure that deposition does not increase to the point of a target load exceedance.

The Regional Lagrangian Acid Deposition (RELAD) model (Cheng and Angle 1996; Cheng et al. 1995, 1997) has been used to estimate the amount of acid deposition in Alberta. There are no grid cells currently receiving acid deposition in excess of their assigned critical or target loads. However, soils in some parts of the province may be sensitive to levels of acid deposition less than the monitoring load (0.17 kmol H^+ ha⁻¹ yr⁻¹) for sensitive ecosystems.

1.2 OBJECTIVES

As stipulated in the Alberta Acid Deposition Management Framework, Alberta Environment is responsible for conducting an evaluation of the acid deposition monitoring data in the province, as well as conducting an evaluation of receptor sensitivity. The present study addresses the latter of these two requirements for the Edmonton East map sheet (NTS 83H East Half).

The objective of this receptor sensitivity study is to provide an estimate of the critical load for the soil types and water bodies present in the Edmonton East study area. On the basis of these critical loads estimates, a recommendation regarding the sensitivity of the study area as a whole is provided.

1.3 EDMONTON EAST GRID CELL STUDY AREA

The study area is the East Half of Map Sheet 83H, located in central Alberta. The boundaries are:

112°W longitude - east side 53°N latitude - south side 113°W longitude - west side, and 54°N latitude - north side

This area is herein referred to as the Edmonton East study area. Landmarks and/or towns located within the grid cell are Elk Island National Park, Bruderheim near the northwest edge, Vegreville near the central east edge, Camrose in the southwest corner, Beaverhill and Whitford Lakes, and other towns such as Tofield, Ryley, Holden, Mundare, Chipman, Lamont, Andrew and Willingdon. Expected areas with Sensitive and Moderate Sensitive soils are the Brunisolic soils north of Bruderheim, northwest of Andrew, and in the far northeast part of the study area. Expected areas with Moderate Sensitive soils are the Luvisolic soils in the Cooking Lake moraine area, including much of Elk Island National Park, and the sandy Chernozemic soils located mainly in the northern half of the study area. The areal extent of the study area is approximately 7,386 square kilometres.

1.4 APPROACH TO CRITICAL LOAD DETERMINATION

In order to determine the appropriate critical load for the study area, it is necessary to determine the soil types and land uses, to chemically analyze samples collected from the various soil types and water bodies present within the study area, and to estimate the site-specific critical load for each sample using a mathematical receptor model. The approach follows the critical loads determination for the Provost-Esther area reported by Turchenek and Abboud (2001) and for the Edmonton West area (Abboud and Turchenek, 2009).

Critical loads are essentially a measure of the buffering capacity of the system. The buffering capacity can be altered by processes other than deposition of acidic substances from the atmosphere. Agricultural and range management practices may have a large impact on soil chemistry and, therefore, make it difficult to assess the relatively small impacts of acid deposition on soils used for agriculture (crop production) or for livestock grazing. For this reason, the emphasis of this project was on soil and water systems that are not, or are minimally, affected by intensive farming and/or range management practices.

The study included a number of components as follows:

- Compilation of available data on soil types, land uses and aquatic systems within the defined area, and generation of a map showing this information.
- Collection of samples of soil and water to determine the critical load for each soil type/land use/aquatic unit.
- Laboratory analysis of the soil and water samples to obtain model input data.
- Using the ARC and Steady State Mass Balance models, and the laboratory data, to estimate the critical load for each sample.
- Extrapolate the site data to provide an estimate of the critical load for each soil type/land use/aquatic unit, expressed in terms of acidification sensitivity categories.
- Development of a map showing the soil acidification sensitivity categories.

2.0 CRITICAL LOADS AND APPROACHES TO THEIR DERIVATION

2.1 CRITICAL LOAD DEFINITION

The term 'critical load' is defined in Alberta as 'the highest load that will not cause chemical changes leading to long-term harmful effects on the most sensitive ecological systems' (Clean Air Strategic Alliance and Alberta Environment 1999). The critical load represents the level of sustained deposition of a substance that will not cause long-term harmful change to an ecosystem. It is thus a property of the ecosystem. The concept of critical loads has been adopted in various countries, especially those of the European Union, as a method for development and implementation of control strategies for air pollutants. Critical load approaches and mapping programs are most extensively developed in Europe, and are described in publications by Downing et al. (1993), Task Force on Mapping (1996), and Posch et al. (1995, 1997, 2003). The applicability of critical loads in Alberta has been discussed in Maynard (1996) and Schindler (1996). Based upon these two reports, critical loads have become the foundation of Alberta's Acid Deposition Management Framework (Clean Air Strategic Alliance and Alberta Environment 1999).

2.2 CRITICAL CHEMICAL CRITERIA AND CRITICAL CHEMICAL VALUES

The process to establish critical loads depends upon the selection of critical chemical criteria. For soils, these criteria are chemical parameters such as pH, base saturation, aluminum (AI) concentration in soil solution, base cation (BC) concentration in soil solution, and the ratio of BC to AI concentrations. Any or all of these may be selected, and critical loads based upon the inputs chosen may be derived. For water the process is similar, with acid neutralizing capacity (ANC) being the most common critical chemical criterion used.

For each critical chemical criterion, critical chemical values must be established (Sverdrup et al. 1990). Critical chemical values are frequently referred to as thresholds. The criteria selected for this study and the rationale for each selection, and the critical chemical values (thresholds) assigned to each criterion, are discussed below.

2.2.1 Soil pH

Soil pH is defined as the pH of a solution in equilibrium with soil. It is determined by means of a glass, quinhydrone, or other suitable electrode or indicator usually using distilled water or a salt solution at a specified soil-solution ratio. Various methods can be used to measure soil pH; those particularly relevant in acid deposition impact evaluations are as follows:

- pH (paste) a soil sample is made into a paste with distilled water, and the pH measured by insertion of an electrode into the paste;
- pH(CaCl₂) a soil sample is mixed in 0.01M CaCl₂ at a 1:2 soil:solution ratio (w:v), and the pH is measured with a glass electrode dipped into the solution;
- pH(H₂O) a saturated paste of soil in water is filtered, and the pH of the filtrate is measured with a glass electrode; and,

pH(solution) - soil solution is extracted in *situ*, and the pH of the solution is measured with a glass electrode.

Theoretically, the pH(solution) measure provides the most realistic indication of the pH environment of plant roots. However, pH(solution) is the most difficult to obtain due to the need for *in situ* extraction equipment and due to the time required to obtain sample for the pH measurement.

The closest estimates of the pH of solution *in situ*, particularly for soils having low soluble ion content, as reflected by low electrical conductivity are provided by $pH(H_2O)$ and pH(solution) (Hendershot et al. 1993). However, accuracy and reproducibility by these methods are difficult to attain because of various factors that can affect the measurement, including soil:solution ratio, position of the measuring electrode, drying of soil, CO_2 concentration, and others. The value obtained may thus not reflect the actual pH of soil solution; however, close estimates of the pH in the root environment can be obtained by controlling some factors, particularly the soil:water ratio (e.g., 1:2 weight:volume).

The pH of soil sample suspended in 0.01 M CaCl₂ solution at a fixed soil:solution ratio is a commonly used method to characterize soil pH. This method has several advantages over pH(H₂O), among them being reproducibility even with dried soil samples. The salt solution generally results in a pH value about 0.5 units lower than that determined in water. Thus, it underestimates the soil solution pH, although it has also been considered to more accurately estimate the pH at the surfaces of soil particles because the weak salt solution simulates the soil electrolyte concentration adjacent to these surfaces. pH(CaCl₂) expresses a relationship between hydrogen and other cations in the soil solution (Bache 1980). Thus, it is responsive to changes in the concentrations of base cations relative to hydrogen, and as such can be useful in monitoring because it would decrease as base cations are lost from soils. Miewes et al. (1986) also noted that pH(CaCl₂) is the more appropriate pH measure for characterizing the buffer range of a soil. Measurement of pH(CaCl₂) and pH(H₂O) measures are most commonly used in research and reported in the literature. Different soil acidification models use different pH measures. Consequently, it is important to indicate which measure is used.

Ulrich et al. (1984) suggested that a soil $pH(H_2O)$ of 4.0 to 4.2 posed a high risk of damage to forest ecosystems, and that there was some risk at $pH(H_2O)$ values greater than 4.2. Low soil pH is typical of forest soils, but is relatively uncommon in grassland soils. Chernozemic soil pH values are typically in the range of 5.6 to 7.7 (Turchenek et al. 1987). Soils in the range of pH(H₂O) 5.6 to 6.0 are sufficiently acidic to cause serious loss in yields of most crops in Alberta (Penney et al. 1977; Hoyt et al. 1981). Turchenek and Abboud (2001), in deterring critical loads for the predominantly Chernozemic soils of the Esther area, suggested that the critical chemical value for pH (4.0 to 4.2) for forest soils is not appropriate for application to grassland soils. Furthermore, the typical range in Chernozemic soil pH values would also suggest that the forest soil criteria are not appropriate for Chernozemic soils under native grassland. Because pH values below 5.6 represent the lower limit of pH values associated with Chernozemic soils (and grassland soils in general), and a reduction in pH below 5.6 could trigger changes in

microbiological and plant species composition, the critical chemical value for $pH(H_2O)$ of pH 5.6 is applied. This is equivalent to a $pH(CaCl_2)$ of about 5.0.

2.2.2 Calcium to Aluminum and Base Cation to Aluminum Ratios

Different threshold levels of Al^{3+} related to plant health have been suggested (Bloom and Grigal 1985; Ulrich et al. 1984; Levine and Ciolkosz 1988); however, Cronan and Grigal (1995) indicated that although total concentration of Al in soil solution might appear to be the most straightforward index of potential Al toxicity to plants, this measure usually fails to be closely related to plant health. This may be due to the differential toxicity of the various Al species and to the ameliorative effects of other ions in solution. Reported Al toxicity thresholds for trees have a wide range, from <40 μ mol L⁻¹ to >3,000 μ mol L⁻¹. However, toxicity has been shown within a much narrower range in terms of the Ca:Al molar ratio (range of 0.2 to 2.5), and risk thresholds are therefore indicated in terms of this latter measure.

Cronan and Grigal (1995) reviewed Ca:AI ratios and other properties as indicators of stress in forest ecosystems and suggested a multiple assessment approach for determining the probability of suffering AI stress. The suggested threshold Ca:AI molar ratio of 1 is commonly applied in setting critical loads for forest soils in European countries (Warfvinge and Sverdrup 1992; de Vries 1993). Little information is available with respect to the significance of Ca:AI ratios in grassland soils, although the same critical chemical value (Ca:AI of 1) has been applied to various types of ecosystems in critical load determinations in Europe (Posch et al. 1997). In some countries, the BC:AI ratio is applied instead of Ca:AI because of work showing that BC:AI correlates more strongly with plant root or shoot damage than Ca:AI. The term 'BC' in this expression refers to the sum of the molar concentrations of the cations Ca, Mg and K.

Sverdrup and Warfvinge (1993) presented a data compilation from the literature showing response curves of growth of seedlings of various tree and ground vegetation species in relation to the BC:AI ratio. The BC:AI ratios at which growth of various grass species was negatively affected ranged widely from 0.3 to 300. Of the species listed, only Kentucky bluegrass (*Poa pratensis*), an introduced species, is found in Alberta grasslands. This species is listed as having a critical BC:AI ratio of 250. Some grasses of the same genus as those found in Alberta (*Festuca, Bromus, Agrostis*), and some *Carices*, have ratios ranging from 1 to 45. Only species of the *Poa* genus have BC:AI ratios of 250 or greater, while the maximum ratio for all other species is 45. Sensitivity of species of the *Festuca, Poa* and *Bromus* genus to pH and AI has also been found by Edmeades et al. (1991).

In the absence of research specific to grasslands in western Canada, it is difficult to select an appropriate BC:AI ratio that would be protective of all species. The ratio applied to forest soils of 1.0 appears to be low for grass species. The median value for the range of grasses reported by Sverdrup and Warfvinge (1993) is about 10. A critical value of 45 had previously been selected in a study of sensitivity of soils in the Provost-Esther grid cell (Turchenek and Abboud, 2001). Most of the soils examined in the Edmonton East grid cell occur in the northern part of the study area where Chernozems are transitional to Brunisolic and Luvisolic forested soils. A transitional BC:AI ratio might, therefore, be more appropriate for these soils. The grass species median range of 10 (Sverdrup and Warfvinge 1993) is thus suggested as the critical chemical value for

Chernozemic soils in the Edmonton East grid cell. However, for purposes of comparison with this suggested critical chemical value for BC:AI, the examination of critical loads in this study includes derivations of critical loads using ratios of 1, 10, 45 and 250.

2.2.3 Base Saturation Percentage

Soil percent base saturation was identified by Cronan and Grigal (1995) and by Miewes et al. (1986) as important in evaluating potential acidification stress on forest ecosystems. While there are various methods of measuring base saturation, the method relevant to threshold limits is based on percent of 'effective cation exchange capacity'. Effective cation exchange capacity (CEC) is defined as the CEC that occurs at field pH, as opposed to CEC measured at a specified pH (i.e., using a pH buffered extractant). Effective CEC is measured by extraction of exchangeable cations using a neutral, unbuffered saturating solution such as NaCl, KCl, BaCl₂ or NH₄Cl. The effective CEC quantifies the number of negatively charged sites with which cations are associated; the major cations in most soils are Ca, Mg, K, Na, Al, Fe, Mn and H. Thus;

$$CEC = Ca+Mg+K+Na+Al+Fe+Mn+H$$
 (expressed as cmoles charge per kg) (1)

Base Sat % =
$$(Ca+Mg+K+Na) \times 100 / (Ca+Mg+K+Na+Al+Fe+Mn+H)$$
 (2)

Ca, Mg, K and Na are categorized as basic cations because the reaction between an exchangeable cation and free H⁺ derived from dissociation of water results in generation of hydroxyl (OH⁻). Al, Fe and Mn, on the other hand are categorized as acidic cations, as they react and tie up OH⁻ from H₂O, resulting in release of an equivalent amount H⁺ (McBride 1994).

The measurement of CEC and base saturation according to equations (1) and (2) rely on measurement of each of the individual cations. An independent measure of CEC can also be obtained. When unbuffered NH_4CI , or other neutral salt solution, is passed through a soil sample, NH_4^+ displaces the exchangeable cations. The NH_4^+ on the exchange complex is then replaced by Na by passing a NaCl solution through the sample, and the amount of NH_4^+ is measured, the quantity of NH_4^+ being equal to the CEC. Base saturation is then calculated as:

Base Sat % =
$$(Ca+Mg+K+Na) \times 100 / (CEC)$$
 (3)

Base saturation can also be calculated from an independent estimate of the portion of the exchange attributable to acid cations (AI, Fe, Mn and H). This measure is referred to as the Exchangeable Titrateable Acidity (ETA). Base saturation is then calculated as:

Base Sat % =
$$(Ca+Mg+K+Na) \times 100 / (Ca+Mg+K+Na) + ETA$$
 (4)

All of the above approaches theoretically provide the same base saturation value, although they seldom do so in practice. Different methods are applied in different institutions and countries. The protocol of the UNECE International Cooperative Programme on Integrated Monitoring (UNECE Convention on Long-Range Transboundary Air Pollution, 2006) applies methodology according to equation (4) above, although the other approaches are used in other programs (e.g., Miewes et al. 1986). Cation exchange capacity values applied in dynamic modelling of

critical loads in Europe are based on measurement in a solution buffered at pH 6.5 (Task Force on Modelling and Mapping, 2004). Thus, methodologies differ between monitoring and modelling applications, and it is important that the specific methods be specified.

For forest ecosystems, a threshold base saturation reduction to a level of 5% (a critical chemical value of 5% base saturation) was suggested by Ulrich et al. (1984), while a reduction to 15% was recommended as a threshold by Cronan and Grigal (1995) on the basis of work by Cronan and Schofield (1990). These threshold values refer to base saturation calculations based on 'effective cation exchange capacity'; that is, cations measured in an extract from a soil sample equilibrated with a neutral salt solution rather than a buffered solution (i.e., Equation 1). A base saturation value of 10%, based on neutral salt exchangeable cation determination, is commonly applied as a critical value of soil chemistry effects.

Low base saturation is a characteristic of forest soils, and forest soils typically have relatively low pH values. Grassland soils, however, are characterized by relatively high base saturation and pH values. Chernozemic soils are the most common grassland soils, with Solonetzic and Vertisolic soils being common associates. A Chernozemic 'A' horizon is diagnostic for the Chernozemic Order of soils in Canada (Soil Classification Working Group 1998). Among the criteria associated with a Chernozemic A horizon is a base saturation greater than 80% and dominance of exchangeable calcium on the exchange complex (other criteria apply to Chernozemic soils, but they are not associated with acidification). This 80% base saturation level is based on measurement by the 'neutral salt' method.

On the Canadian Prairies there is a gradual change in the nature of surface soil horizons from grassland soils in the south to forested soils in the north, where leached (Ae) horizons become more prevalent. A leached Ae horizon is indicative of loss of base cations and decreased pH in this horizon. There is thus a relationship between vegetation and the type of surface soil, the implication being that vegetation changes as pH and base saturation decrease. Climate, however, is another major factor that prevents grasslands on the dry prairie from converting to forest vegetation if they become acidified. It might be hypothesized, however, that prairie vegetation assemblages would change in response to acidification, such that more acid tolerant species may become more prevalent. On this basis, therefore, a base saturation of 80% appears to be an applicable threshold limit for acidification of grassland soils.

As noted above, the 80% base saturation criterion for Chernozemic soils is based on measurement by the 'neutral salt' method, and an equivalent value based on a 'buffered' CEC measurement is not provided in the Canadian System of Soil Taxonomy (Soil Classification Working Group 1998). In 'Soil Taxonomy', the system of soil classification applied in the United States, a base saturation of at least 50%, determined by the ammonium acetate buffered method, is a criterion for definition of a mollic epipedon. The mollic epipedon is similar in definition to the Chernozemic A horizon, which is diagnostic of Chernozemic soils in the Canadian system of soil classification. Consequently, since the ARC model utilizes the base saturation based on a pH 7.0 buffered extraction procedure, a base saturation of 50% based on an ammonium acetate measurement of CEC could be adopted as a critical chemical value for Chernozemic soils. This is applied together with the $pH(H_2O)$ criterion (Section 2.2.1) as an

indicator of acidification effects in this report. Further discussion about application of base saturation to critical load determination is provided in Section 5.3.6.

2.2.4 Acid Neutralizing Capacity (ANC) of Aquatic Systems

Acid neutralizing capacity (ANC) is the ability of a solution to neutralize inputs of strong acid to a pre-selected equivalence. It is calculated as:

$$[ANC] = [BC] - [AN] = [HCO_3^{-1}] + [A^{-1}] - [H^{+1}] - [AI^{n+1}]$$
(5)

where, [BC] is the base cation concentration, [AN] is the strong acid anion concentration, $[HCO^{3-}]$ is the bicarbonate concentration, [A⁻] is the organic anion concentration and $[AI^{n+}]$ is the sum of all inorganic Al ions. A threshold (critical chemical value) for ANC of 20 μ mol L⁻¹ has been applied in Scandinavia as a critical chemical value for fish in surface waters (Henriksen et al. 1990), although different ANC values specific to different receptors have also been suggested (Henriksen et al. 1995). The threshold is applied in models used to determine critical loads for surface water bodies; e.g., the Steady State Water Chemistry model and the First-Order Acidity Balance model (Task Force on Mapping 1996).

2.2.5 Summary

The threshold or critical chemical value refers to the value of a critical chemical criterion or combination of criteria (e.g. ratios) above or below which no harmful response in a biological indicator is expected occur. The critical chemical values pertinent to forest and grassland soils and to surface waters that are used in this study are given in Table 1.

Critical Chemical Criteria (Indicators)	Critical Chemical Values (Thresholds)						
Soils							
pH(CaCl ₂) - Forest Soils ²	3.5						
pH(H ₂ O) - Forest Soils ²	4.2						
pH(CaCl ₂) - Grassland Soils ^Y	5.0						
pH(H ₂ O) - Grassland Soils ^Y	5.6						
Base saturation percentage - Forest Soils ^x	<10% of effective CEC						
Base saturation percentage - Grassland Soils Y	<80% of effective CEC						
	1.0 (50% risk)						
BC:Al ratio - Forest Soils [×]	0.5 (75% risk)						
	0.2 (95-100% risk)						
BC:Al ratio - Grassland Soils Y	10						
Surface water							
ANC ^W	20 µmol L ⁻¹						

Table 1. Proposed Indicators and Thresholds of Stress in Forest and GrasslandEcosystems

² After Ulrich et al. (1984)

^Y After Sverdrup and Warfvinge (1993)

^x After Cronan and Grigal (1995)

^w After Task Force on Mapping (1996)

2.3 EMPIRICAL METHOD FOR DERIVATION OF CRITICAL LOADS

Empirical methods of critical load derivation are based mainly on observation of responses of ecosystem components to acid deposition. In the case of soils, it has been suggested that a basic principle underlying a critical load is that the total input of hydrogen ions to the soil must not exceed the alkalinity produced by the weathering of soil minerals (Nilsson 1986). At a workshop in Skokloster, Sweden, it was concluded that the rate of chemical weathering is the single most important factor governing the soils ability to buffer incoming acidity, and therefore critical loads, for forest soils (Nilsson and Grennfeldt 1988).

This mineralogical approach (the Skokloster approach) was adopted with some modifications for critical load determination of soils in the U.K. (Hornung et al. 1995). Texture, drainage, soil thickness and other factors were considered in deriving critical loads in the U.K. Details of the application of this mineralogical approach to the Edmonton East study area are presented in Section 5.1.

2.4 USE OF MODELS TO DERIVE CRITICAL LOADS

Numerous models have been developed to examine soil acidification and to derive critical loads. Modelling approaches comprise two main categories referred to as 'steady-state methods' and 'dynamic modelling'. Within each category, there are varying degrees of sophistication ranging from simple calculations to complex mathematical constructs. The most complex are integrated forest soil models that link soil processes to other processes such as vegetation growth, hydrology and nutrient cycling.

Steady-state models calculate deposition levels that avoid harmful effects in ecosystems in steady-state (Task Force on Mapping 1996). Processes such as cation exchange and sulphate adsorption have a finite time scale and therefore cannot be included in steady-state models. Therefore, steady-state models are mainly used for calculation of critical loads over very long periods of time. Two types of steady-state models have been developed for soils. One-layer models, such as the Steady State Mass Balance (SSMB) model consider the soil as a single layer, whereas the multi-layer models consider chemical conditions in different soil layers or horizons. The one-layer SSMB model has been the most commonly applied tool for derivation of critical loads of soils in Europe (Task Force on Mapping 1996).

Dynamic models are a family of more complex models that use various calculations to simulate changes in soil solution or water chemistry due to acid deposition over time. Examples are the MAGIC, SAFE, VSD and SMART models, which have been developed in Europe (UNECE Convention on Long-Range Transboundary Air Pollution, 2006), and the ARC model, applied in this report. Calculations of critical loads using these models is not as straightforward as with steady-state models because of the temporal aspect; i.e., it is necessary to determine the acceptability or non-acceptability of chemical changes in soils or waters in relation to a predetermined period of time. Another reason for non-usage is the need for much data required to run some of the dynamic models. Consequently, dynamic models have not been used to a great extent in determining critical loads. However, these models are useful in scenario analysis; i.e., for assessing effects of given deposition levels over a selected period of time, and for determining the effects of different emission abatement strategies.

Dynamic models are used to calculate the acidification process for an ecosystem through time. Dynamic models, as compared to steady-state models, require more input data of which several parameters are more difficult to obtain. Since an assessment of the time periods involved in acidification responses and recovery from acidification can be made with these models, they are the best tools available for addressing time-dependent scenarios and the impact of episodic events on ecosystems. Several of the models listed here are research tools, and are not available for evaluation and application in Alberta at the present time.

The gradual change with time in the acidification state of the system in response to some change in deposition is calculated with dynamic models. Critical loads can be calculated from different deposition scenarios, and the results can be compared to the critical chemical values (thresholds) for several different critical chemical criteria (e.g., ion exchange, weathering of soil minerals, uptake and cycling of base cations and nitrogen by plants, and soil solution equilibrium chemistry) in the system simultaneously. They use integrated mass balances for substances and differential equations for the rates of different processes. The time-dependent scenarios are obtained by numerical integration of the model subroutines advancing in small time-steps.

Various assumptions are made in the equations within the dynamic models. It is generally assumed that the CEC is constant over time and that a certain ion exchange equilibrium applies (Gapon or Gaines-Thomas), and aluminum is assumed to be in continuous equilibrium with a mineral of the same composition as gibbsite (de Vries 1991). Some models assume sulphate adsorption to be negligible or at steady-state, while others have sulphate adsorption as a major process.

Some soil models are subroutines of more complex models used for impact studies and critical load determinations for aquatic systems. Sverdrup et al. (1990) suggested that several models be examined before choosing a model for soil evaluations and critical load calculations. The models differ somewhat in their basic principles, and have different limitations connected to their use and to the interpretation of their results. Such factors must be carefully studied before a model is chosen for a specific type of system. Brief descriptions of some currently used dynamic models are provided below.

The availability of data is a major consideration in determining the method to be used for critical load determination. This factor generally limits the methods to empirical methods or to steady-state and the simpler dynamic modelling approaches.

2.4.1 Steady-State Mass Balance (SSMB) Model

The Steady-State Mass Balance model is calculated manually and can be used for quick evaluation of scenarios involving relatively higher and lower levels of acid deposition and neutralizing capacities to arrive at critical loads. This is a one-layer model wherein only a specified thickness of the soil profile can be considered. Details are presented in Section 5.2.1

Critical load determination by the SSMB model is directly dependent on the weathering rate, which is the major long-term source of alkalinity that neutralizes acidity in the soil system and the major source of base cations for replacing those removed by leaching. Thus, confidence in the critical load determined by this method depends on the level of confidence in the model input value for the weathering rate. Most estimations of weathering rate are based on correlations of experimentally determined weathering rates with soil type, mineralogy, base cation content or texture. Others are based on computations using soil mineralogy, wherein quantitative data for the complete suite of minerals present in a soil are required. The approach has been widely used in Europe to provide a weathering term for input into the SSMB equation (Task Force on Mapping 1996).

Another approach to estimating weathering is based on an estimation of mineralogy from total chemical analysis of soil by use of the UPPSALA model which performs a stepwise allocation of elements (Ca, Mg etc.) to different soil minerals. Minimal data needed by the UPPSALA model for converting elemental contents to mineralogy are levels of total Na, K, Ca, Mg, P, Al, Si and Fe (Sverdrup 1990).

The SSMB approach is applicable in Alberta in terms of the three criteria of simplicity, availability and applicability in critical load derivation. Critical loads can easily be calculated for an individual soil, or a large number of computations can be made within a spreadsheet. The SSMB model was applied to the Edmonton East study area using a weathering rate estimated from information in the literature as described in Section 6.2.1.

2.4.2 Alberta Research Council (ARC) Model

The ARC model is derived from Bloom and Grigal (1985) and incorporates empirical relationships for cation exchange and pH based on Alberta soil properties. The model has been described by Abboud and Turchenek (1990) and Turchenek et al. (1993), and is described in part in Section 5 of this report.

2.5 SURFACE WATER ACIDIFICATION MODELS

The determination of critical loads of acidity to surface waters was an initial objective in determining critical loads in the Edmonton East study area. However, the high salinity of lakes in the region results in very low acidification sensitivity. Thus, there was no effort made in determining the critical loads of surface waters in this area.

2.6 MODELS USED TO DERIVE CRITICAL LOADS FOR SOILS IN THE EDMONTON EAST STUDY AREA

The SSMB and ARC models were previously applied to determination of critical loads in the Provost-Esther area (Turchenek and Abboud, 2001) and the Edmonton West study area. An empirical method was also applied, in which critical loads were based on the Skokloster method.

3.0 DATA ACQUISITION AND COMPILATION METHODS

3.1 BASELINE SOIL INFORMATION

Information about the distribution and properties of soils in the Edmonton East study area is available from soil survey reports and from the AGRASID database (Alberta Soil Information Centre, 2007). The AGRASID database provides soil survey coverage for the agricultural regions of Alberta, along with descriptions of soil series, including typical soil chemical attributes. Soil distribution is presented in the database within a hierarchical framework based on the national ecological framework for Canada (Ecological Stratification Working Group 1995). Details of ecological stratification and soil properties in the study area are presented in Sections 5.1 and 5.2.

The Edmonton East study area is within the Prairies Ecozone. An Ecozone is an area that is representative of large and very generalized ecological units characterized by interactive and adjusting abiotic and biotic factors.

An Ecoregion is a part of an Ecozone characterized by distinctive ecological responses to climate as expressed by the development of vegetation, soil, water, fauna, etc. (Ecological Stratification Working Group 1995). The study area occurs within the Aspen Parkland ecoregion, with the northern edges bordering the Boreal Transition Ecoregion. The Aspen Parkland is characterized by predominance of Black Chernozemic soils, with inclusions of Gleysolic and Solonetzic soils. These soils transition to the predominantly Luvisolic soils of the Boreal Plains ecoregion, which are associated with Brunisolic soils where materials are coarse textured. Gleysols occupy poorly drained depressions, and Organic (peat) soils occur increasingly toward the northern part of the area.

An Ecodistrict is a subdivision of an Ecoregion in the ecological land classification hierarchy. It is characterized by distinct assemblages of landform, relief, surficial geologic material, soil, water bodies, vegetation and land uses (Ecological Stratification Working Group 1995). The soil mapping system in Alberta further subdivides Ecodistricts into Land Systems. A Land System is defined as a subdivision of an Ecodistrict that is recognized and separated by differences in one or more of general pattern of land surface form, surficial geologic materials, amount of lakes or wetlands, or general soil pattern. All Land Systems within one Ecodistrict have the same general climate for agriculture, but differences in microclimatic pattern can be recognized. Soil Landscapes are subdivisions of Land Systems that display a consistent and recognizable pattern of distribution of soils and landscape elements (Alberta Soil Information Centre, 2007).

Soil types as identified at the Land System level were applied in developing a sampling protocol and critical loads map of the study area. Analysis of soil types at the Soil Landscape level of mapping would prove to be unwieldy due to the large number of delineations within one grid cell. Land Systems provide information at a lower level of detail, but at a somewhat greater level than that of the land units that form the basis of soil sensitivity mapping by Holowaychuk and Fessenden (1987). Consequently, Land System information was considered to be a practical basis for refining the previous soil sensitivity mapping and for calculating critical loads.

3.2 INITIAL ACID SENSITIVITY RATING

Each Land System is characterized by an assemblage of dominant and subdominant soil series. The extent of each series was estimated from the attribute information provided in AGRASID. These were then allocated an acid sensitivity rating based on base loss, acidification, aluminum solubilization and overall sensitivity ratings using soil pH and cation exchange capacity as the major criteria (Holowaychuk and Fessenden 1987). The ratings were developed for the top 20 cm of soil. However, soil chemical data reported in soil survey reports are based on one or very few sampled profiles, and it is difficult to fully rely on these data for sensitivity classification. Of the soil attributes described in soil surveys reports, texture would be considered as one that is frequently and reliably estimated in the field. Cation exchange capacity is strongly related to texture because of its dependence on the clay content of the soil. Thus, instead of applying the Holowaychuk and Fessenden (1987) sensitivity classification using chemistry data only, soils in the Edmonton East study area were assigned preliminary sensitivity ratings on the basis of texture as well. Soils of sand or loamy sand texture were characterized as being Sensitive to acid deposition. Soils of sandy loam texture were assigned a Moderate sensitivity rating. Luvisols were mainly assigned a Moderate rating because the topsoils (A horizons) commonly have sandy loam textures, even though the underlying material is fine textured. All Chernozemic soils of texture finer than sandy loam (including fine and very fine sandy loam) were assigned a Low sensitivity rating.

The sensitivity rating allocation to Land Systems provided information about coverage of all potentially acid sensitive soils. This provided a framework for representative sampling of soils for the critical loads evaluation. Agricultural soils and native/range soils of Low sensitivity were excluded from the evaluation because (1) acid deposition management is to be based on the extent of sensitive soils affected (Clean Air Strategic Alliance and Alberta Environment 1999), and (2) these soils are generally under cultivation and subject to various management practices, particularly fertilization, which confound any evaluations of acidification due to atmospheric deposition. In the Edmonton East study area, native rangelands were included, although these consist mainly of soils under native forest. Open forage and range areas have generally had tree cover removed and consist of non-native species. Although these lands have been cultivated and possibly fertilized, soil samples were taken from some sites in order to examine their potential sensitivity to acid deposition.

The sensitivity ratings were re-evaluated upon completion of the critical load determinations, with allocations to sensitivity classes based on pH, base saturation percentage and base cation to aluminum ratio. These were compared with the criteria of Holowaychuk and Fessenden (1987), and a revised soil distribution and acid sensitivity map was produced for the study area.

3.3 LAND USE INFORMATION

As indicated previously, forage and crop lands are subjected to various practices such as fertilization and manure application, and these would complicate evaluations in relation to atmospheric acid deposition. Additionally, soils under cultivation are generally soils that have higher nutrient content and buffering capacity (base cations), and are therefore the least sensitive soils within any given area. Land use information was therefore required, in addition to

soil and landscape information, to enable planning of a sampling program, and more importantly, to enable calculation of the areal extents of soils of different acid sensitivity.

Land use information was available from the Prairie Farm Rehabilitation Administration (PFRA) which had undertaken mapping for the purpose of verifying applications under the Western Grain Transition Payments Program (WGTPP). The WGTPP map was based on analysis of satellite images acquired from 1993 to 1995, and land cover was allocated to one of eleven classes:

- 1. Cultivated crop land land that is annually seeded or under summer fallow;
- 2. Forage (hay) land that is in perennial forage for hay or silage production (dominantly alfalfa);
- 3. Grasslands land that is in perennial grasses and herbaceous species for grazing use including native range, seeded tame pasture, abandoned farm areas and other non-cultivated uses (ditches, riparian areas, etc.);
- 4. Shrubs land that has perennial woody shrub coverage;
- 5. Trees hardwoods, mixed woods, recent burns and cutovers;
- 6. Wetlands intermittent water bodies, areas that have semi-permanent or permanent wetland vegetation, including fens, bogs, swamps, sloughs, marshes, etc.;
- 7. Water permanent water bodies including lakes, rivers, irrigation canals;
- 8. Non-agricultural lands land that is dominantly in a non-vegetative or non-agricultural land use, including farmsteads, roads, cities, towns, open pit mines, industrial sites, etc.;
- 9. Clouds and shadow;
- 10. Mud, sand and/or saline areas; and,
- 11. Unclassified area areas outside of the study area.

Areas classed as Shrubs or Trees (categories 4 and 5 above) were selected from the WGTPP digital database and superimposed on the initial soil and soil sensitivity map, the development of which is described in Section 4.1 above. All other land was regarded as tilled land, although minor areas of disturbed lands (in addition to urban arrears) occurred as well. Spatial information about water bodies was then taken from a separate digital layer in the database to produce a combined soil/land use/surface water map.

There can be uncertainty in the classification of certain types of land in the PFRA land classification. Moreover, the imagery that the classification was based on is now dated. This is nevertheless the most readily available land cover database. An inherent assumption in the sensitivity analysis herein is that this land cover information is more or less accurate, and that it is adequate for deriving statistics for areas of soils with different sensitivity ratings.

3.4 BASELINE SURFACE WATER INFORMATION

Information about the distribution and extent of surface water bodies in the Edmonton East study area was derived form the WGTPP information as indicated in Section 4.2. Detailed information about the areas of wetlands within the ecosystems of Alberta is available in 'Ecodistricts of Alberta: Summary of Biophysical Attributes' (Strong and Thompson 1995) and in 'Characterization of Wetlands in the Settled Area of Alberta' (Strong 1993). Information about shallow water bodies is included in the latter compilation, but lakes are not included. Water quality information is available in the form of a digital database maintained by Alberta

Environment. The database presents values for pH, alkalinity, total dissolved solids and calcium for more than 1,000 Alberta Lakes, with information about additional lakes added on an ongoing basis (Saffran and Trew 1996).

3.5 Soil Sampling

The goals established for soil sampling to meet the needs of critical load determination were to obtain soil samples of the LFH and the top 25 cm of mineral topsoil at a minimum of 25 sites from the Sensitive and Moderate acidification sensitivity areas in the study region.

The initial soil and soil sensitivity rating (Section 3.2) resulted in identification of 12 Land Systems that have a component of potentially Sensitive or 'Sensitive plus Moderate' soils, and one additional Land System in which soils of potential Moderate sensitivity to acidic deposition occur. These areas varied in size. Sampling within the Edmonton region was logistically challenging in terms of obtaining permission to enter lands, and finding suitable areas for sampling within relatively densely populated areas such as acreage developments. Consequently, locations categorized as natural areas, parks and other crown lands were targeted for collection of samples. Some sites outside the boundaries of the study area were selected, provided they were located within land systems that extended into the Edmonton East study area.

Soil samples were taken by excavating a small pit to at least 50 cm depth and taking a volume of about 2 L each of LFH (forest floor) and 0-25 cm horizons. In most instances, the 0-25 sampling layer occurred entirely within the A horizon. In some case where the A horizon was thinner than 25 cm, a portion of the B horizon to the 25 cm depth was included in the sample. The samples were collected in October and early November, 2006.

3.6 WATER SAMPLING

No water samples were taken from the water bodies within the study area. There are few lakes within the Edmonton East map sheet. Five lakes in the provincial database (Beaverhill, Hastings, Miquelon, Tawayik and Islet) had alkalinity values ranging from 168 to 1,627 mg L⁻¹, and were thus considered to be of Low sensitivity to acidification. Whitford, Cucumber, Limestone, Dusty and Demay Lakes, along with other small, unnamed lakes occur mainly within areas with saline subsoils, and these were not considered to be potentially sensitive to acidification.

3.7 SOIL ANALYSES

Soil samples were analyzed for various properties as follows:

pH(CaCl₂): By potentiometric measurement in a 1:2 (w:v) solid-to-liquid mixture of soil in 0.01 M CaCl₂. (Method 7 (ii) in Kalra and Maynard (1991). The soil-to-solution ratio for litter (LFH) material was 1:4.

 $pH(H_2O)$: By potentiometric measurement in a saturated paste (Method 7 (i), Kalra and Maynard (1991).

Calcium Carbonate Equivalent: Carbonates were dissolved by reaction with HCl and the evolved CO_2 was measured manometrically as described in Method 10 of Kalra and Maynard (1991), with the exclusion of timed measurements for differentiation of calcite and dolomite.

Total Carbon, Nitrogen and Sulphur: By combustion and automated detection using a Leco C-N-S unit. The total-C obtained was corrected for carbonate-C, if present, to obtain total organic-C. Samples with pH <7 were assumed to contain no carbonate-C, and, therefore, total-C equals total organic-C in these cases.

Cation Exchange Capacity (Neutral Salt): By 1.0 M NH₄Cl extractant (unbuffered), and measurement of NH₄⁺ by distillation. The method is described in Method 15 (ii) in Kalra and Maynard (1991). The distillation step differed in that NH_4^+ is not displaced with Na, but the whole sample was distilled to determine content of adsorbed NH_4^+ .

Cation Exchange Capacity (Buffered): By 1.0 M ammonium acetate extractant buffered at pH 7, and measurement of NH_4^+ by distillation. The method was applied as described in Procedure 3.3.2 in McKeague (1978), except that NH_4^+ was not displaced with Na, and the whole sample was distilled to determine the content of adsorbed NH_4^+ .

Exchangeable ions: By Inductively Coupled Plasma (ICP) Atomic Emission Spectroscopy of the unbuffered CEC extract. Ions included in the ICP scan were Ca, Mg, Na, K, Fe, Mn and Al. Exchangeable H was estimated from the pH difference between the unbuffered NH₄CI extractant before and after extraction.

Electrical Conductivity and Soluble Salts: By measurement of electrical conductivity and ions in the aqueous extract from a saturated paste of a soil sample (Method 8(i), Kalra and Maynard (1991). EC and pH were measured in the extract. A portion of the extract was filtered using a 0.45-µm micropore filter, and a full ICP elemental scan, including S, was conducted on the extract.

3.8 METEOROLOGICAL DATA

Precipitation data were obtained directly from the Atmospheric Environment Service, Environment Canada, for the years 1990 - 2000 (Environment Canada, 2006). This was supplemented by data from Canadian Climate Normals, 1961 - 1990 (Environment Canada, Atmospheric Environment Service, 1993). Data were obtained for the meteorological station at the Edmonton International Airport.

3.9 **PRECIPITATION SURPLUS**

Some models use the term 'precipitation minus potential evapotranspiration' to obtain an approximation of the amount of deep percolation of soil moisture, or to approximate total precipitation surplus including runoff. Potential evapotranspiration exceeds precipitation in the study area, however, the depth of soil profile development suggests that water penetrates to about 0.8 metres in sandy soils in the study area. A soil depth of 25 cm was applied in

determining acidification with models, this being the depth within which the majority of plant roots occur. Therefore, the amount of water percolating beyond the surface 25 cm zone was calculated.

Daily precipitation data for the months of April to October, inclusive, were obtained for the years 1990 to 1995. The amounts of precipitation retained by the soil on a daily basis was estimated by assuming a field capacity of 16.7 mm per 25 cm, this being based on an available water content of 80 mm per 1.2 metres for sandy soils in the study area (Tajek et al. 1989). The daily evapotranspiration rates were subtracted from this amount. Actual monthly evapotranspiration rates were obtained from Bothe and Abraham (1993). These rates were as follows: April, 2 mm d⁻¹; May, 4 mm d⁻¹; June, 6 mm d⁻¹; July, 7 mm d⁻¹; August, 6 mm d⁻¹; September, 4 mm d⁻¹; and, October, 2 mm d⁻¹. All winter snowfall was assumed to percolate into the soil, and evapotranspiration was assumed to be zero for this period.

The amount of percolation beyond the 25 cm zone varied widely among the five years. The average amounts calculated for the northern, central and southern parts of the study area were 107, 92 and 78 mm per year, respectively.

The difference between the precipitation and the precipitation surplus represents the proportion of the precipitation that reacts with the upper 25 cm soil layer. Another implication of the precipitation surplus concept is that the products of any reactions within the top 25 cm of the soil are carried down the profile; that is, base cations may be lost from the upper layer.

While most roots are assumed to occur in the top 25 cm, a proportion occurs at some depth in the profile and take up nutrients as well as water. Thus, it is possible that upward movement of nutrients through these deep roots would add nutrients to the upper soil layers, which would serve to counteract the effects of acidification on plants whose roots occur in the top 25 cm of the soil. However, it is difficult to estimate the amount of upward nutrient transport by deep roots. It is considered that this is a minor process within the ecosystem, and therefore, this amount was not estimated and it was assumed for modelling purposes that no nutrient return occurs by this mechanism.

3.10 ACID DEPOSITION DATA

Acid deposition data were obtained from province-wide estimates of deposition by Cheng et al. (1997). For the ARC model, the Potential Acid Input (PAI) was applied. The PAI reported by Cheng et al. (1997) for the Edmonton East map sheet was 0.15 to 0.20 kmol H⁺ ha⁻¹ yr⁻¹, and more recently the estimate was 0.10 to 0.17 kmol H⁺ ha⁻¹ yr⁻¹ (WBK & Associates Inc., 2006). The upper number in this range (0.20 kmol H⁺ ha⁻¹ yr⁻¹) was applied in models. This rate is equivalent to 0.3 kmol H⁺ ha⁻¹ yr⁻¹ of SO_x, NO_x and NH_x deposition, partially neutralized by 0.1 kmol H⁺ ha⁻¹ yr⁻¹ of base cation deposition.

3.11 Other Data Requirements

Other model data inputs consisted of constants, coefficients, soil analytical data or soil parameters obtained from the literature, or they have been derived for Alberta soils (see Section 5).

3.12 Map Compilation

A soil map of the Edmonton East study area was developed from the AGRASID soils database and PFRA land cover databases as described in Section 4.2. Only information at the Land System level was used, as this is an appropriate level of detail for depicting the distribution of soil types and their sensitivity to acid deposition. Additionally, it provided a suitable level of stratification for planning a soil sampling program.

Digital files for base map information as well as land use data were obtained from the PFRA-WGTPP data base (Section 4.3). The base map files were registered to UTM Zone 12, NAD '83 coordinates. This coordinate system was maintained throughout all digital processing and formed the basis for geographic referencing of the final map products.

Delineations of Land Systems from the AGRASID database were linked to the base information using ARC/VIEW. The data were exported to ARC/INFO for topological construction, attribute linkage and map product output, the latter including incorporation of a soil sensitivity legend.

Site-Specific Critical Loads of Acid Deposition on Soils in the Edmonton 83H East Map Sheet, Alberta

4.0 OVERVIEW OF SOILS AND SURFACE WATERS IN THE STUDY AREA

4.1 ECOLOGICAL STRATIFICATION

The most detailed level of mapping in the AGRASID database is the Soil Landscape unit. A Soil Landscape is a subdivision of a Land System that displays a consistent and recognizable pattern of distribution of soils and landscape elements (Alberta Soil Information Centre, 2007). As indicated in Section 4.1, the Soil Landscape mapping unit was considered as being too detailed for application in this project, and the Land System was applied instead.

A map of Land Systems in the study area is presented in the back pocket of this report. A legend describes characteristics of the Land Systems in terms of parent geologic materials, landscapes and soil types. The Land Systems are also described in Table 2.

4.2 SOIL CLASSIFICATION

The study area occurs within the Aspen Parkland ecoregion, with the northern edges bordering the Boreal Transition Ecoregion. The Aspen Parkland is characterized by predominance of Black Chernozemic soils, with inclusions of Gleysolic and Solonetzic soils. These soils transition to the predominantly Luvisolic soils of the Boreal Plains ecoregion, which are associated with Brunisolic soils where materials are coarse textured. Gleysols occupy poorly drained depressions, and Organic (peat) soils occur increasingly toward the northern part of the area.

Individual soil types within Soil Landscapes are identified at the Soil Series level of the Canadian System of Soil Classification (Soil Classification Working Group 1998). A soil series is a category (or level) in the Canadian system of soil classification. It is the basic unit of soil classification, and consists of soils that are essentially alike in all major profile characteristics except the surface texture. Naming of Soil Series is based on the Alberta Soils Names File (Generation 3) User's Handbook and Soil Correlation Area (SCA) Map of Alberta (2006) (Alberta Soil Information Centre 2007). Soil series within the study area are listed in Table 3.

4.3 SOIL AND LAND COVER MAP

Soil types, land use and distribution of surface water bodies are shown on the map 'Land Systems, Land Cover and Soil Sensitivity to Acid Inputs in the Edmonton East map sheet' (back pocket). The surficial materials consist mainly of glacial till, glaciolacustrine, glaciofluvial and fluvioeolian deposits ('fluvioeolian' refers to a complex of glaciofluvial deposits with eolian deposits occurring as blankets and dunes). Landscapes range from undulating to hummocky. A legend accompanying the map indicates the dominant and minor soil series within each Land System, along with the parent materials and landscape features.

Land cover in the study area was categorized as cultivated, grassland, shrubland or treed land. The distribution and extent of these land cover types is indicated in the Land System map (back pocket).

Land System Symbol	Land System Name	Land System Description	Major Soils	Minor Soils
05.00.05	North Saskatchewan River Valley	Landscape is valley bottom with some numerous water bodies and confined floodplain. Regosols developed on undifferentiated material. Minor soils include Chernozems and Gleysols. Significant eroded soils present.	ZERzbl ^z	PED ZGW
05.3d.09	Partridge Plain	Landscape is undulating. Black Chernozems developed on medium textured till. Minor soils include Gleysols.	AGS	RLV ZGW
05.3d.11	Pointe-Aux-Pins Plain	Landscape is undulating. Black Chernozems developed on fine textured water-laid sediments	MMO	AGS LOM
05.3d.27	Ferlow Plain	Landscape is hummocky. Black Chernozems developed on medium textured till. Minor soils include Gleysols.	AGS	RLV ZGW
05.4a.02	Lonestar Plain	Landscape is undulating. Thin Black Solonetz developed on medium textured till. Minor soils include Chernozems.	KLM	HER DYD
05.4a.03	Irys Plain	Landscape is undulating. Thin Black Solonetz developed on medium textured till and till over softrock. Minor soils include Gleysols.	KLM SHS	ZGW DYD
05.4a.04	Bruce Plain	Landscape is undulating. Thin Black Solonetz developed on medium textured till. Minor soils include Gleysols.	KLM	DYD ZGW
05.4a.05	Daysland Plain	Landscape is undulating. Thin Black Chernozems developed on medium textured till. Minor soils include Gleysols.	HER	EOR FMN
05.4a.06	Bawlf Plain	Landscape is undulating. Thin Black Solonetz developed on medium textured till. Minor soils include Chernozems and Gleysols.	KLM	HER ZGW
05.4a.14	Little Beaver Plain	Landscape is undulating. Black Chernozems developed on medium textured till. Minor soils include Solonetz and Gleysols.	NRM	CMO ZGW
05.4a.15	Ryley Plain	Landscape is undulating. Black Solonetz developed on medium textured till. Minor soils include Chernozems and Gleysols.	СМО	NRM ZGW
05.4a.16	Beaverhill Lake	Large water body. Minor soils include Gleysols.	ZWA	ZGW
05.4a.17	Chipman Plain	Landscape is undulating. Black Solonetz and Black Chernozems developed on medium textured till. Minor soils include Gleysols.	CMO AGS	ZGW KVG
05.4a.19	Katchemut Upland	Landscape is hummocky. Black Chernozems developed on medium textured till. Minor soils include Gleysols.	AGS	NRM ZGW
05.5b.04	Royal Park Plain	Landscape is undulating. Black Chernozems developed on medium textured till and medium textured material over medium textured till. Minor soils include Gleysols.	AGS HBM	ZGW
05.5b.05	Inland Plain	Landscape is undulating with some valleys with confined floodplain. Black Chernozems and Black Solonetz developed on medium textured till. Minor soils include Gleysols.	NRM AGS CMO	ZGW
05.5b.06	Vegreville Plain	Landscape is undulating with some floodplain. Black Solonetz and Black Chernozems developed on medium textured till. Minor soils include Gleysols.	CMO AGS	ZGW ZSZzbl
05.5b.08	Whitford Plain	Landscape is undulating with some level and numerous water bodies. Black Chernozems developed on medium textured till. Minor soils include coarse and fine textured soils.	AGS	UKT NVR
05.5b.09	Norma Plain	Landscape is undulating. Black Chernozems and Black Solonetz developed on medium textured till. Minor soils include coarse textured soils.	AGS CMO	MDR PHS
05.5b.12	Watt Lake Plain	Landscape is undulating with some valleys with confined floodplain. Black Chernozems and Black Solonetz developed on medium textured till. Minor soils include Gleysols.	AGS CMO	HYL NRM
05.5b.13	Hairy Hill Plain	Landscape is hummocky. Dark Gray Chernozems developed on medium textured till. Minor soils include coarse textured soils.	RLV	RDW POK
05.5b.14:	Hilliard Plain	Landscape is undulating with some valley bottom. Black Chernozems developed on medium textured till. Minor soils include Solonetz and Gleysols.	AGS	CMO ZGW
05.5b.15:	Kahwin Plain	Landscape is undulating. Black Chernozems developed on medium textured till. Minor soils include Gleysols.	AGS	POK ZGW
05.6.01	Islet Upland	Landscape is hummocky. Dark Gray and Dark Gray Luvisols developed on medium textured till. Minor soils include Gleysols, Chernozems and fine textured soils.	COA UCS	ZGW MCO

Table 2. Description of Land Systems in the Edmonton East Grid Cell

² See Table 3 for Major and Minor soil series descriptions.

Land System Symbol	Land System Name	Major Soils	Minor Soils	
06.2a.01	Musidora Upland	Landscape is undulating. Dark Gray and Black Chernozems developed on coarse textured sediments. Minor soils include Brunisols and Organic.	HLW PHS	PRM ZOR
06.2a.05	Redwater Plain	Landscape is undulating with some duned. Black Chernozems developed on coarse textured sediments. Minor soils include Brunisols.	MDR	PRM PHS
06.2a.11	Eldorena Plain	Landscape is undulating with some duned. Black Chernozems and Brunisols developed on coarse textured sediments.	PHS PRM	MNTaa MDR
06.2b.05	Edward Upland	Landscape is hummocky and duned. Brunisols and Black Chernozems developed on coarse textured sediments. Minor soils include Luvisols and Organic.	PRM RDW	UCS ZOR
06.2b.08	Redclay Plain	Landscape is undulating with some inclined < 10% exposed bedrock. Black Chernozems developed on medium textured till and medium textured water-laid sediments.	AGS POK	RMY RLV
06.2b.12	Delph Upland	Landscape is hummocky. Black Chernozems and Dark Gray Luvisols developed on medium textured till. Minor soils include Gleysols.	AGS UCS	RLV ZGW
06.2b.17	Pakan Plain	Landscape is undulating. Black Chernozems developed on medium textured water-laid sediments. Minor soils include Solonetz.	POK	HBM KVG
06.2c.25	Landscape is undulating. Dark Gray Luvisols and Dark Gray			

Table 2. Description of Land Systems in the Edmonton East Grid Cell (concluded)

Symbol	Series	Drainage	Calcar	-	PM1 Texture	РМ1 Туре	PM1 Texture	PM2 Type	Soil Subgroup	Subgroup Modifier
AGS	Angus Ridge	W	М	Ν	MF	TILL	-	-	E.BL	
CMO	Camrose	W	М	М	MF	TILL	-	-	BL.SS	
COA	Cooking Lake	W	М	Ν	MF	TILL	-	-	O.GL	
DYD	Daysland	W	М	W	MF	TILL	-	-	BL.SO	
EOR	Elnora	W	М	Ν	MF	TILL	-	-	O.BL	
FMN	Foreman	Р	W	М	MF	TILL	-	-	SZ.HG	
HBM	Hobbema	W	М	Ν	ME	GLLC	MF	TILL	E.BL	
HER	Heisler	W	М	W	MF	TILL	-	-	SZ.BL	
HLW	Helliwell	W	W	Ν	VC	GLFL	-	-	O.DG	
HYL	Hairy Hill	Р	М	М	MF	TILL	-	-	R.HG	CRSA
KHW	Kehiwin	W	М	Ν	MF	TILL	-	-	D.GL	
KLM	Killam	W	М	М	MF	TILL	-	-	BL.SS	
KVG	Kavanagh	MW	W	W	MF	SRFS	-	-	BL.SS	
LOM	Looma	W	W	N	VF	GLLC	MF	TILL	O.DG	
MCO	Mico	MW	М	N	VF	GLLC	-	-	O.DG	
MDR	Mundare	W	W	Ν	VC	GLFL	-	-	O.BL	
MMO	Malmo	W	W	Ν	FI	GLLC	-	-	E.BL	
MNTaa	Manatokan-aa	VP	N	N	0	FNPT	MC	GLFL	T.M	
NRM	Norma	W	М	Ν	MF	TILL	-	-	SZ.BL	
NVR	Navarre	I	W	N	FI	GLLC	-	-	GL.BL	
PED	Penhold	W	М	N	ME	GLLC	-	-	O.BL	
PHS	Peace Hills	W	W	Ν	MC	GLFL	-	-	O.BL	
POK	Ponoka	W	М	N	ME	GLLC	-	-	E.BL	
PRM	Primula	R	N	N	VC	GLFL	-	-	E.EB	
RDW	Redwater	W	W	Ν	MC	GLFL	-	-	O.DG	
RLV	Rolly View	W	М	N	MF	TILL	-	-	O.DG	
RMY	Rimbey	W	М	N	ME	GLLC	-	-	O.DG	
SDN	Spedden	W	М	Ν	MF	TILL	-	-	O.DG	
SHS	Shonts	W	W	W	MF	TILL	MF	SRFS	BL.SS	
UCS	Uncas	W	М	Ν	MF	TILL	-	-	D.GL	
UKT	Ukalta	W	М	Ν	MC	GLFL	MF	TILL	O.BL	
ZWA	Misc. Water	VP	-	-	-	-	-	-		
ZGW	Misc. Gleysol	Р	-	-	-	UNDM	-	-	O.HG	
ZSZzbl	Misc. Solonetzic-zbl	W	-	-	-	UNDM	-	-	BL.SS	
ZOR	Misc. Organic	VP	-	-	-	UNDO	-	-	TY.M	
ZERzbl	Misc. Eroded-zbl	W	-	-	-	UNDM	-	-	R.BL	

Table 3. Soil Series in the Edmonton East Map Sheet

Source: AGRASID 3.0. Alberta Soil Information Centre: http://www1.agric.gov.ab.ca/\$department/deptdocs.nsf/all/sag6903)

Abbreviations:

Drainage: VR - very rapid; R - rapid; W - well; MW - moderately well; I - imperfect; P - poor; VP - very poor.

Calc (calcareousness) and Salinity: N - non; W - weak; M - moderate

PM1 (upper parent material), PM2 (lower parent material):

PM Texture: VC - very coarse; C - coarse; GRVC - gravelly very coarse; MC - moderately coarse; GRMC gravelly moderately coarse; ME - medium; MF - moderately fine; FI - fine; VF - very fine

PM Type: TILL - glacial till, or morainal; GLFL - glaciofluvial; FLUV - fluvial; FLEO - fluvioeolian; GLLC -

glaciolacustrine; SRFS - soft rock; FNPT - fen peat; SPPT - sphagnum peat; UNDM - undetermined

Soil Subgroup: Defined below (Table based on the Canadian System of Soil classification

Subgroup modifier: CRSA - carbonated and saline

Order	Great Group	Subgroups
Brunisolic - Sufficient development to exclude from the Regosolic order, but lack degrees or kinds of development specified for other orders.	<u>Eutric Brunisol</u> - Ah<10 cm, pH>5.5 <u>Dystric Brunisol</u> - Ah<10 cm, pH<5.5	E.EB - Eluviated Eutric Brunisol E.DYB - Eluviated Dystric Brunisol
Regosolic - Development too weak to meet requirements of any other Order.	Regosol - Ah<10 cm, Bm absent or <5 cm <u>Humic Regosol</u> - Ah≤10 cm, Bm absent or <5 cm	(Not in above table)
Chernozemic - Surface horizons darkened by accumulation of organic matter from decomposition of grassland vegetation.	<u>Black Chernozem</u> - Black Ah, semiarid climate <u>Dark Gray Chernozem</u> - Dark Gray Ah, semiarid climate	O.BL - Orthic Black E.BL - Eluviated Black SZ.BL - Solonetzic Black O.DG - Orthic Dark Gray
Gleysolic - Features indicative of periodic or prolonged water saturation, and reducing conditions - mottling and gleying.	<u>Humic Gleysol</u> - Ah≥10 cm, no Bt <u>Gleysol</u> - Ah≤10 cm, no Bt <u>Luvic Gleysol</u> - Has a Btg, usually has an Ahe or an Aeg	R.HG – Rego Humic Gleysol SZ.HG - Solonetzic Humic Gleysol Various Gleysol subgroups occur in ZGW units (Table 3), including: O.LG - Orthic Luvic Gleysol HU.LG - Humic Luvic Gleysol O.G - Orthic Gleysol
Luvisolic - Light coloured eluvial horizons - Ae; illuvial B horizons of silicate clay accumulation - Bt; developed under forest vegetation.	<u>Gray Luvisol</u> - May or may not have Ah, has Ae and Bt, usually MAST ≤8 degrees Celsius ^Y	O.GL - Orthic Gray Luvisol D.GL - Dark Gray Luvisol GL.GL - Gleyed Gray Luvisol GLD.GL - Gleyed Dark Gray Luvisol BR.GL - Brunisolic Gray Luvisol
Solonetzic - Has Solonetzic B horizon - Bn or Bnt - columnar or prismatic structure, hard to extremely hard when dry, exchangeable Ca/Na≤10.	<u>Solonetz</u> - Lack a continuous Ae≥2 cm <u>Solodized Solonetz</u> - Ae≥2 cm, intact columnar Bnt or Bn <u>Solod</u> - Ae≥2 cm, distinct AB or BA (disintegrating Bnt)	B.SZ - Black Solonetz BL.SS - Black Solodized Solonetz BL.SO -Black Solod
Organic - Composed dominantly of organic materials; most are water saturated for prolonged periods.	<u>Mesisol</u> - Dominantly mesic <u>Fibrisol</u> - Dominantly fibric	T.F Terric Fibrisol T.M Terric Mesisol TF.M - Terric Fibric Mesisol TM.F - Terric Mesic Fibrisol TY.F - Typic Fibrisol M.F - Mesic Fibrisol TY.M - Typic Mesisol F.M - Fibric Mesisol

Table 4. Soil Orders and Great Groups in the Edmonton East Map Sheet

^z Source: Soil Classification Working Group (1998).

^Y MAST = mean annual soil temperature.

4.4 DESCRIPTION OF SAMPLED SOILS

Locations and descriptions of soils sampled in the study area are presented in Appendix A. Analytical data for the soils are presented in Appendix B.

4.5 SURFACE WATERS

The largest water bodies in the area are Beaverhill and Whitford Lakes, both of which are located in the undulating plains region east of the Cooking Lake upland. Larger lakes within the Cooking Lake upland are Hastings, Miquelon, Tawayik and Astotin Lakes; the latter two are located in Elk Island National Park. Two named lakes occur in the southern part of the

Edmonton East map sheet, namely Demay and Dusty Lakes. Other named lakes occurring in the northern half of the study area are Cucumber, Limestone, and Soda Lakes.

Some of the lakes in the study area have data reported in the Alberta Environment Online Lake Water Quality Data database (Table 5). The sensitivity of these lakes to acidification was based on the criteria provided by Palmer and Trew (1987), which is based on the total alkalinity of the lake water. The criteria are:

•	High sensitivity	Alkalinity 0-4 mg L ⁻¹
•	Moderate sensitivity	Alkalinity 5-8 mg L ⁻¹
•	Moderate - Low Sensitivity	Alkalinity 9-25 mg L ⁻¹
•	Low sensitivity	Alkalinity 26-40 mg L ⁻¹

• Least sensitive Alkalinity >40 mg L^{-1}

Based on the water quality of lakes reported in the Edmonton East study area, all lakes have alkalinity levels that greatly exceed levels in the above criteria and can be regarded as least sensitive. Palmer and Trew (1987) did not categorize any lakes in the Edmonton area as being more sensitive than the 'Least sensitive'.

Other lakes were not sampled as part of this study. Examination of soil maps and direct observations indicated that all lakes east of the Cooking Lake upland are associated with saline soils, and the waters in therefore be characterized by high alkalinity. A number of small lakes occur within Elk Island National Park. The relatively high alkalinity of Tawayik Lake was considered to be representative of these, and all are regarded as being of Least Sensitive to acidification.

Lake	Location	рН	Ca (mg L ⁻¹)	Alkalinity (mg L ⁻¹ CaCO ₃)	TDS (mg L ⁻¹)	EC (μS cm⁻¹)	Acidification Sensitivity
Beaverhill	35-51-18-4	8.8	32	391	822	390	Low
Hastings	20-51-20-4	9.1	33	253	551	747	Low
Miquelon	29-49-20-4	9.3	7	2,230	7,140	10,100	Low
Tawayik	21-53-20-4	na	na	326	654	390	Low
Islet	2-52-20-4	8.5	36	173	173	342	Low

 Table 5. Water Chemistry of Lakes in the Edmonton East Study Area

It was concluded that all, or almost all, surface waters in the Edmonton East can be categorized as having Low sensitivity to acidifying inputs. Derivation of critical loads was not, therefore, carried out for any of the surface waters in the study area.

5.0 CRITICAL LOAD DETERMINATIONS FOR SOILS

5.1 EMPIRICAL METHOD

The empirical method as adapted in the UK from the Skokloster approach (Section 2.3) was applied to soils in the Edmonton East study area. The application of this approach begins with allocation of a soil to a particular sensitivity and critical load class based on the dominant minerals in the soil (Table 6). This scheme places soils dominated by clay minerals in the second class. However, the exchange capacity and exchangeable cations carried by clay minerals are not taken into account, and placing a clayey soil into Class 2 was not considered as being appropriate (Hornung et al. 1995). Therefore, a particle size classification was developed for modifying the initial mineralogically-based classes (Table 7). In addition to the soil textural modifiers, various other factors were considered in determining the final classification ratings for different soil types (Table 8). As an example, a soil overlying quartzite bedrock would be allocated to Class 1 in the Skokloster classification system. However, if the soil was poorly drained and loamy-sand in texture, it would be allocated to Class 2.

neralogical Classification and Critical Loads for Skokloster Classification ^z	r Soils (0-0.5 m) According to

Class	Dominant Weatherable Minerals	Critical Load (kmol H⁺ ha⁻¹ yr⁻¹)
1	Quartz, rutile, anatase, kaolinite, gibbsite, orthoclase	< 0.2
2	Muscovite, plagioclase, illite, montmorillonite, vermiculite	0.2 - 0.5
3	Amphibole, chlorite, biotite, epidote, glaucophane	0.5 - 1.0
4	Olivine, garnets, pyroxenes, epidote	1.0 - 2.0
5	Carbonates	> 2.0

^Z After Nilsson and Grennfeldt (1988) and Sverdrup and Warfvinge (1988)

Table 7. Allocation to Skokloster Material Class Based on Particle S	Size Class ¹
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Particle size class	Soil material class
Sand, loamy sand, sandy loam, (sandy) silt loam	Class 2
Clay loam, sandy clay loam, silt loam	Class 3
Clay, silty clay, sandy clay	Class 4

After Hornung et al. (1995).

Factor	Decrease	Increase
Precipitation	High	Low
Vegetation	Coniferous forest	Deciduous forest
Elevation slope	High	Low
Soil texture	See Table 7	See Table 7
Soil drainage	Free	Impeded
Soil/till depth	Shallow	Thick
Sulphate adsorption capacity	Low	High
Base cation deposition	Low	High

Table 8. Factors Causing a Decrease or Increase in Critical Loads of Acidity for Soils²

² After Nilsson and Grennfeldt (1988) and Hornung et al. (1995).

The combination of mineralogical and particle size classes of sand to sandy loam soils in the study area would result in allocation to a critical load category of 0.2-0.5 kmol ha⁻¹ yr⁻¹ (class 2). In the UK approach, the critical load is either increased or decreased, depending on various modifying factors, as indicated above. The factors of low precipitation, low elevation, and thick soil would increase the critical load. However, the factors of free drainage and low sulphate adsorption capacity serve to reduce the critical load. The base cation deposition rate is another modifying factor; the level in the study area, however, is of intermediate magnitude (Cheng et al. 1997) and therefore has little impact on the overall rating. The factors more or less balance each other, and we therefore deduce that very sandy soils (sand, loamy sand) likely have a critical load in the range of 0.2 - 0.5 kmol ha⁻¹ yr⁻¹. This would apply particularly to soils with low organic matter content. The classification for sandy loam soils is likely in the upper part of the range, and possibly in the 0.5-1.0 kmol ha⁻¹ yr⁻¹ range. Allocation of soil units using this empirical method leads to the assignment of critical loads in the Edmonton East study area as presented in Table 9.

Texture	Soil Series	Critical Load
Very coarse	Primula, Nestow, Mundare, Helliwell	0.2-0.5 kmol ha ⁻¹ yr ⁻¹
Moderately coarse	Peace Hills, Redwater, Ukalta	0.5-1.0 kmol ha ⁻¹ yr ⁻¹
Medium to moderately fine	Series on till	1.0-2.0 kmol ha ⁻¹ yr ⁻¹
Fine	Glaciolacustrine clays	>2.0 kmol ha ⁻¹ yr ⁻¹

Table 9. Critical Loads in the Edmonton East Area Based on the Empirical Method

5.2 STEADY STATE MASS BALANCE METHOD

5.2.1 Model Description

The Steady State Mass Balance (SSMB) model considers the soil as consisting of one compartment equal to the thickness of the root zone (generally 30-50 cm or more in forest soils), and calculates critical loads in relation to critical chemical values related to element concentrations leaching from the root zone. The calculation of critical loads using the SSMB model is based on a balance of sources of acidity against sinks for acidity and sources of alkalinity, and uses a formulation of the charge balance of ions in the soil leachate.

Sverdrup and de Vries (1994) and de Vries (1991) provided description and derivation of the model, and the model as applied in Europe is described in UBA (2004). The method was applied in calculating critical loads of acid deposition for forest soils in eastern Canada and most recently in Manitoba and Saskatchewan forest areas (Aherne and Watmough, 2006). Some model assumptions in an earlier approach, applied in the study of critical loads in the Provost-Esther area in Alberta (Turchenek and Abboud, 2001), differ from those applied in the recent Canadian studies. The method applied herein is as described in the Canadian studies, with emphasis on the Manitoba/Saskatchewan study, from which some of the input data were obtained. The critical load of acidity arising from sulphur, CL(S), and from nitrogen, CL(N), is described by the following equation:

$$CL(S) + CL(N) = BC_{dep} - Cl_{dep} + BC_{w} - BC_{u} + N_{i} + N_{u} + N_{de} - Alk_{le(crit)}$$
(5)

where, BC_{dep} is base cation deposition ($BC = Ca^{2+} + Mg^{2+} + K^{+} + Na^{+}$), Cl_{dep} is Cl^{-} deposition, BC_w is base cation weathering, N_i is nitrogen immobilization, N_u is nitrogen uptake by vegetation, and N_{de} is denitrification. Alk_{le(crit)}, the critical alkalinity leaching (also referred to as critical acid neutralizing capacity) is estimated from the critical base cation to aluminum ratio (BC:AI) in the soil solution that leaches through the system along with a term that describes the gibbsite equilibrium, which is assumed to control the Al concentration.

Critical load has also been defined in terms of potential acidity as:

$$CL(Ac_{pot}) = BC_w - BC_u + N_i + N_u + N_{de} - Alk_{le(crit)}$$
(6)

 BC_{dep} nor CI_{dep} are not considered in the definition because they are ecosystem properties and can change over time (UBA, 2004). The nitrogen terms have been assumed to be nil or very close to nil in applications to Canadian soils, and removal of base cations is generally not considered. In forest soils, base cations would be removed by harvesting; in grassland situations, this term would be minimal as the main export of cations would be via livestock. Since cations are not removed, the BC_u term is considered to be nil. The critical load potential acidity is then defined as:

$$CL(Ac_{pot}) = BC_{w} - Alk_{le(crit)}$$
(7)

Critical 'Alkalinity leaching' (Alk_{le(crit)}) can be defined in terms of soil acidity as follows:

$$AIk_{le(crit)} = -AI_{le(crit)} - H_{le(crit)} = -Q \cdot ([AI]_{crit} + [H]_{crit})$$
(8)

Q is the precipitation surplus, or water leaving the root zone (m³/ha/yr), and the square brackets denote concentrations (in eq/m³).

The relationship between AI and H is defined by the gibbsite equilibrium:

$$[AI] = K_{qibb} \bullet [H]^3 \text{ or } [H] = ([AI]/K_{qibb})^{1/3}$$
(9)

The Alk_{le(crit)} term is then defined as,

$$Alk_{le(crit)} = -Q^{2/3} \cdot \{1.5 \cdot (BC_{dep} + BC_w - BC_u)/((BC:AI)_{crit} \cdot K_{gibb})\}^{1/3} - 1.5 \cdot (BC_{dep} + BC_w - BC_u)/(BC:AI)_{crit})$$
(10)

where Q is the precipitation surplus, or water leaving the root zone ($m^3 ha^{-1} yr^{-1}$). Values for the parameters are presented in the following section. K_{gibb} is the gibbsite equilibrium constant.

The incorporation of these relationships in the CL expression (equation 8) provides the SSMB equation for critical load of acidity in mol ha⁻¹ yr⁻¹, as follows:

$$CL(Ac_{pot}) = BC_{w} + \{1.5 \cdot (Bc_{w} + BC_{dep} - BC_{u})/((BC:AI)_{crit} \cdot Kgibb)\}^{1/3} \cdot Q^{2/3} + 1.5 \cdot (BC_{w} + BC_{dep} - BC_{u})/(BC:AI)_{crit})$$
(11)

Q is the precipitation surplus, or water leaving the root zone (m³ ha⁻¹ yr⁻¹). Values for the parameters are presented in the following section.

The full derivation of the equation and the explanation of factors used in the ANC_{le(crit)} term can be found in UBA (2004).

5.2.2 Data for SSMB Critical Load Calculations

Precipitation Surplus (Q)

Q is calculated as the precipitation minus the sum of interception evaporation by vegetation, the actual soil evaporation and the actual transpiration (water uptake) in the root zone. The precipitation surplus term is discussed in Section 5.3.1.3. The SSMB calculations were carried out for a 75 cm soil layer, which is consistent with the approach elsewhere in Canada (Aherne and Watmough, 2006). For the ARC model (Section 5.3.1.3), the estimate is 780 m³ ha⁻¹ yr⁻¹ for percolation out of the 25 cm soil layer, and the estimate for percolation below the 75 cm depth is 200 m³ ha⁻¹ yr⁻¹.

Gibbsite Equilibrium Constant (Kgibb)

The value of K_{gibb} depends on soil type and the organic matter content. The value for soils with low organic matter ranges from 300 - 3,000 m⁶ mol⁻² (UBA, 2004). $K_{gibb} = 300 \text{ m}^6 \text{ mol}_c^{-2}$ was applied in modelling for the Edmonton East grid cell.

Weathering Rates (BCw)

A number of options for estimating weathering rates are presented by the Task Force on Mapping (1996) and more recently in UBA (2004), and were previously described in detail by Sverdrup and de Vries (1994) and Sverdrup (1990). Application of these methods to data presented by Sverdrup (1990) for sandy soils suggests that the weathering rate is in the range of 0.05 to 0.4 kmol H⁺ ha⁻¹ yr⁻¹ for a 1 metre soil layer, or about 0.01 to 0.1 kmol H⁺ ha⁻¹ yr⁻¹ for a 0.25 metre layer. A value of 0.07 kmol H⁺ ha⁻¹ yr⁻¹ for a 0.25 m layer has been determined for sandy soils in Minnesota by Bloom and Grigal (1985), and was subsequently considered as a suitable approximation for sandy soils in Alberta by Abboud and Turchenek (1990), Turchenek and Abboud (2001) and Turchenek et al. (1994). This rate was therefore applied to soils in the Edmonton East grid cell.

In keeping with SSMB applications in other parts of Canada, a 0.75 m soil layer was applied in modelling. Although most plant roots generally occur within the uppermost soil horizon, the depth of soil exploited by roots can be much deeper. Weathering rates are described further in Section 5.3.1.1. The weathering rate of 0.07 kmol H⁺ ha⁻¹ yr⁻¹ 0.25 m⁻¹ (from Turchenek and Abboud, 2001) was applied in the case of sandy soils. The equivalent 0.75 m weathering rate is 0.21 kmol H⁺ ha⁻¹ yr⁻¹ (210 mol H⁺ ha⁻¹ yr⁻¹ applied in the model). For other textures, the weathering rates were as follows (expressed as mol ha⁻¹ yr⁻¹): loamy sand to sand soil – 300; sandy loam – 450; loam – 750; clay loam – 1,500; and clay or heavy clay – 3,000.

Growth Uptake or Export of Base Cations (Bc_u)

Over a long-term, the net uptake of base cations (BC_u; Ca, Mg and K, with Na excluded) is equal to that stored in vegetative biomass. In the case of grasslands, annual growth (biomass) is returned to the soil each year, and cation storage levels in biomass are considered to be negligible. Nutrients can also be "exported" from soils through livestock grazing and removal of livestock from the land. Little data is available for nutrient removal rates from rangelands by animals. Heady and Child (1994) reported exports of 0.025 to 0.035 kmol H⁺ ha⁻¹ yr⁻¹ of base cations from rangelands in New Mexico. These data suggest that export of nutrients by beef production is low, and rates for northern climates on poor soils would likely be even lower, due to lower stocking rates. Therefore, base cation export by animals is considered as negligible for purposes of deriving critical loads by the SSMB or other methods, and was set to zero in the model runs.

BC:Al Ratio

Base cation to aluminum ratios used in the calculations were 1, 10, 45 and 250 (see Section 2.2.2).

5.2.3 Critical Load Calculations

Critical loads were calculated using the SSMB model (equation 8) for a 0.75 m soil layer. Critical load calculation by the SSMB method was not conducted on the basis of properties of the soil samples, but rather on the basis of weathering rates of broad soil groupings (sand, sand to loamy sand, sandy loam, loam, and clay loam soils) and on regional variation in climate expressed as variation in precipitation surplus. Base cation export was assumed to be negligible.

The results of SSMB calculations (Table 10) showed that critical loads at the BC:Al ratio of 2 are one and a half to two times greater than those at BC:Al ratios of 10. (Note: The expression 'BC' is used heretofore, although it is equivalent to 'Bc' defined above.) However, increasing the BC:Al ratio beyond 10 reduced the critical load only slightly. The lowest critical loads were obtained for the sandy to loamy sand soils.

	Texture	Soil	Major Soil	Crit	ical Load	(kmol ha ^{⁻1}	yr ⁻¹)
Soil Type	Group	Texture Group	Series ^z	BC:AI 1	BC:AI 10	BC:AI 45	BC:AI 250
Eluviated Dystric Brunisol Eluviated Eutric Brunisol	Very Coarse	Sand, Loamy Sand	Nestow Primula	0.6	0.3	0.2	0.2
Eluviated Black Chernozem Dark Gray Chernozem	Very Coarse - Moderately Coarse	Loamy Sand	Helliwell Mundare	0.7	0.4	0.4	0.3
Orthic Black Chernozem Eluviated Black Chernozem Dark Gray Chernozem	Moderately Coarse	Sandy Loam	Peace Hills Redwater Ukalta	1.0	0.6	0.5	0.5
Black Chernozem Eluviated Black Chernozem Dark Gray Chernozem	Medium	Loam, Silt Loam	Ponoka Hobbema	1.6	1.0	0.8	0.8
Orthic Gray Luvisol Dark Gray Luvisol Orthic Black Chernozem Eluviated Black Chernozem Dark Gray Chernozem Black Solodized Solonetz Solonetzic Black Chernozem	Medium, Moderately Fine	Sandy Clay Loam Clay Loam	Cooking Lake Uncas Rolly View Angus Ridge Elnora Heisler Camrose Killam Norma Shonts	2.9	1.8	1.6	1.5
Orthic Black Chernozem Eluviated Black Chernozem Dark Gray Chernozem	Fine, Very Fine	Heavy Clay	Malmo Mico Looma	5.6	3.6	3.2	3.1

 Table 10. Critical Load Calculations by the SSMB Method

² "Major soil" occurring in study area Land Systems, from Table 2.

5.3 ARC MODEL

The ARC model simulates mineral soil chemical processes directly related to acidity and acidification of soils, and predicts the associated soil properties of pH, base saturation, solution Al³⁺ concentration and base cation to aluminum (BC:Al) ratio. The ARC model is described in detail in Turchenek et al. (1993), Abboud and Turchenek (1990) and Abboud et al. (2002). This model is adapted from the Bloom and Grigal (1985) model, with modifications for calculations of acid inputs and acidification processes, method of output of model results, and inclusion of calculations for base cation to aluminum (BC:Al) ratio. These are described in greater detail in the following sections.

5.3.1 Data for Critical Load Determinations

The model requires climatic, soil and acid input data with a provision for varying time period for exposure and a varying time increment for reporting simulation results.

5.3.1.1 Soil Data Inputs

Soil data inputs for the ARC model are as follows:

pH - by the water paste method; if the pH data were reported in a $CaCl_2$ solution (1:2), then the following equations (developed for mineral soils from a correlation of pH values using data from Pauls et al. (1996) were used to transform into a water paste pH:

for LFH horizons: $pH(H_2O) = 0.96 pH(CaCl_2) + 0.55$ R² = 0.989, n= 65 samples (12) for mineral horizons: $pH(H_2O) = 0.94 pH(CaCl_2) + 0.72$ R² = 0.984, n= 130 samples (13)

Cation exchange capacity and exchangeable bases - by the ammonium acetate extraction

Activity coefficients of monovalent, divalent and trivalent ions - activity coefficients for each modelled soil horizon were calculated from the mean values for individual members of that

series.

method.

The activity coefficients (γ_i) were calculated using the Davies equation (Lindsay 1979).

$$Log \gamma_i = -AZ_i^2 \left[\{ I/(1+I^{0.5}) \} - 0.3 I \right]$$
(14)

Where A = 0.509 for water, Z is ion valence and I is ionic strength in moles L^{-1} .

Partial pressure of CO₂ - assumed to be 0.005 atmosphere.

The ionic strengths (I) were calculated from the electrical conductivities of the saturated paste extracts (Lindsay 1979).

$$I = 0.013 EC$$
 (15)

where I is in moles L⁻¹ and electrical conductivity (EC) of the saturated paste extracts in dS m⁻¹.

Initial weathering rates (kmol ha⁻¹ yr⁻¹) for mineral soils - these varied with soil texture as discussed in Abboud et al. (2002) and shown in Table 11 below.

_

Table 11.	Weathering Ra	tes Suggested for Modell	ing Soils of Different Textures ²

Soil Texture	Weathering Rate in 25 cm Surface Soil Layer (kmol ha ⁻¹ yr ⁻¹)
Sand	0.07
Loamy Sand	0.10
Sandy Loam	0.15
Loam, Silt Loam	0.25
Clay Loam, Silty Clay Loam, Sandy Clay Loam	0.50
Clay, Silty Clay	1.00

² From Abboud et al. (2002)

The input data for soil pH, CEC, and sum of bases were weighted mean values for the whole LFH layer (usually less than 25 cm) and the top 25 cm of air-dried mineral soil. The thickness of the soil horizons and the bulk density were applied in computing the means. The calculations were made as previously documented by Turchenek et al. (1993) and Abboud and Turchenek (1990).

5.3.1.2 Acid Deposition Data

The ARC model was applied using a range of PAI values to enable determination of critical loads. The loads used in this modelling exercise were 0.1, 0.2, 0.3, 0.5, 0.7 and 1.0 kmol H⁺ ha⁻¹ yr⁻¹. These values were recommended for model application by the Alberta Environment staff and cover existing PAI values and potential extreme future values encountered in the study area. The PAI values account for both wet and dry forms of acid deposition.

5.3.1.3 Climate Data

Data for precipitation and precipitation surplus as described in Abboud et al. (2002) were applied in the model. Previous applications of the model used a 'precipitation minus potential evapotranspiration' term to determine the amount of precipitation water that percolates beyond the 25 cm layer. This calculation results in a negative value for climates characteristic of central and southern Alberta. The precipitation surplus concept (Abboud et al. 2002) provides a more realistic approximation of the amount of water that is actually evaporated or transpired by accounting for episodes of high precipitation and deep moisture percolation.

5.3.1.4 Time

The model can be executed for any specified length of time, and simulation results can be reported for any specified increment of time within the total simulation period. Predictive soil effects data are of greatest interest in terms of the immediate and near future; i.e., the period during which pollutant emissions can be forecast. It is also of interest, from a soil development point of view, to determine soil responses to acid deposition over very long periods of time since changes in soils occur slowly. Three hundred years was selected for the simulation period. This time frame would not obscure the data for interpretation of short-term effects, yet would provide a longer term view of soil changes.

A one year increment of time between reported values in the simulations was selected. This increment assured that sufficient data points were obtained for determining the trends of pH, base saturation, Al³⁺ levels and BC:Al over time.

5.3.1.5 Effect of Weathering

The weathering (r) of soil minerals is estimated in the model by the function,

$$r = r_0 10^{-0.5(pH-pH_0)}$$
(16)

where r_o and pH_o are the initial conditions (Abboud et al., 2002). The r_o value is based on soil texture as shown in Table 11, and a pH_o of 5.0 was applied in the equation.

5.3.1.6 Summary of Data Inputs

The starting parameters for soils used in simulations are given in Table 12. The taxonomy and some general descriptive features of the soils are indicated along with input data described previously.

 Table 12. Input Data for Soil Acidification Simulations with the ARC Model

System Definition Variables					Soil Data				
Land System	Islet Upland 1	Islet Upland 2	lslet Upland 3	lslet Upland 4	Islet Upland 5	Eldorena/ Redwater Plains	Edward Plain 1	Edward Plain 2	Watt Lake Plain
Sites	13, 16, 18	17	11	12	19	5, 6, 7, 8, BR	3	4	1
Soil Subgroup	Orthic Gray Luvisol	Orthic Gray Luvisol	Dark Gray Luvisol	Dark Gray Luvisol	Orthic Dark Gray Chernozem	Eluviated Dystric Brunisol	Eluviated Dystric Brunisol	Eluviated Dystric Brunisol	Eluviated Black Chernozem
Soil Series	Cooking Lake	Cooking Lake (Acidic)	Uncas (Forage)	Uncas	Helliwell-GR	Nestow	Nestow-TA	Nestow	Mundare (Native Pasture)
Texture 0-25 cm	Sandy Loam	Loam	Sandy Loam	Sandy Loam	Loamy Sand	Sand	Loamy Sand	Sand	Loamy Sand
Precipitation (cm yr ⁻¹)	50	50	50	50	50	50	50	50	50
Litter ET (cm yr ⁻¹)	20	20	19	20	18	14	14	14	0
Perc below 25 cm (cm yr ⁻¹)	14	14	14	14	19	22	22	22	18
Years of Iteration	300	300	300	300	300	300	300	300	300
Increment of Years	1	1	1	1	1	1	1	1	1
PAI (kmol _c H⁺ ha⁻¹ yr⁻¹)	0.1, 0.2, 0.3, 0.5, 0.7, 1.0	0.1, 0.2, 0.3, 0.5, 0.7, 1.0	0.1, 0.2, 0.3, 0.5, 0.7, 1.0	0.1, 0.2, 0.3, 0.5, 0.7, 1.0	0.1, 0.2, 0.3, 0.5, 0.7, 1.0				
LFH	,,	,,	,,	,,	,,	,,	,,	,,	,,
LFH (cm)	10	7	0 (forage)	13	6	2	5	7	0
LFH pH (CaCl ₂)	6.0	5.5	-	6.5	6.2	4.4	5.2	4.8	-
LFH pH (H ₂ O)	6.3	5.8	-	6.8	6.5	4.8	5.5	5.2	-
LFH Bases (kmol _c ha ⁻¹)	78	59	-	130	32	7	50	33	-
LFH CEC (kmol _c ha ⁻¹)	82	91	-	130	33	14	62	50	-
Activity Coefficient of Al1+	0.898	0.906	-	0.890	0.894	0.916	0.903	0.916	-
Activity Coefficient of Al ²⁺	0.650	0.673	-	0.628	0.638	0.705	0.664	0.705	-
Activity Coefficient of Al ³⁺	0.380	0.411	-	0.351	0.380	0.457	0.398	0.455	-
Slope of pH-BSat Equation	3.76	3.76	3.76	3.76	3.76	3.76	3.76	3.76	3.76
Mineral 0-25 cm	•							•	
Mineral Soil pH (CaCl ₂)	4.7	4.1	6.6	5.1	5.2	4.6	5.1	5.0	5.9
Mineral Soil pH (H ₂ O)	5.1	4.6	6.9	5.5	5.6	5.0	5.5	5.4	6.3
Mineral Bases (kmol _c ha ⁻¹)	318	245	442	348	216	57	370	169	690
Mineral CEC (kmol _c ha ⁻¹)	451	716	450	462	336	121	507	268	707
Activity Coefficient of Al1+	0.948	0.961	0.915	0.948	0.935	0.953	0.937	0.946	0.911
Activity Coefficient of Al ²⁺	0.809	0.852	0.703	0.808	0.763	0.827	0.772	0.800	0.687
Activity Coefficient of Al ³⁺	0.621	0.698	0.456	0.620	0.546	0.655	0.565	0.607	0.430
Slope of pH-BSat Equation	2.27	2.27	2.27	2.27	3.38	2.06	2.06	2.06	3.38
CO ₂ Partial Pressure (atm)	0.005	0.005	0.005	0.005	0.005	0.005	0.005	0.005	0.005
Weathering (kmol _c ha ⁻¹ yr ⁻¹)	0.15	0.25	0.15	0.15	0.10	0.07	0.10	0.07	0.10

System Definition Variables					Soil Data				
Land System	Musidora Upland 1	Musidora Upland 2	Musidora Upland 3	Musidora Upland 4	Musidora Upland 5	Whitford Plain		Norma Plain 1	Norma Plain 2
Sites	24	22	25	9, 10, 23, 26	15	2	21	TH1	20
Soil Subgroup	Orthic Dark Gray Chernozem	Orthic Dark Gray Chernozem	Orthic Dark Gray Chernozem	Eluviated Dystric Brunisol	Orthic Black Chernozem	Orthic Black Chernozem	Orthic Dark Gray Chernozem	Orthic Black Chernozem	Orthic Black Chernozem
Soil Series	Helliwell	Helliwell-GR (Forage)	Helliwell (Forage)	Nestow	Peace Hills (Forage)	Ukalta (Native)	Redwater	Mundare (Forage)	Ukalta
Texture 0-25 cm	Loamy Sand	Loamy Sand	Sand	Sand	Sandy Loam	Loamy Sand	Loamy Sand	Loamy Sand	Sandy Loam
Precipitation (cm yr ⁻¹)	50	50	50	50	50	50	50	50	50
Litter ET (cm yr ⁻¹)	18	0	0	14	0	0	18	0	14
Perc below 25 cm (cm yr ⁻¹)	19	23	23	22	18	18	19	23	22
Years of Iteration	300	300	300	300	300	300	300	300	300
Increment of Years	1	1	1	1	1	1	1	1	1
PAI (kmol _c H⁺ ha ⁻¹ yr ⁻¹)	0.1, 0.2, 0.3, 0.5, 0.7, 1.0	0.1, 0.2, 0.3, 0.5, 0.7, 1.0	0.1, 0.2, 0.3, 0.5, 0.7, 1.0	0.1, 0.2, 0.3, 0.5, 0.7, 1.0	0.1, 0.2, 0.3, 0.5, 0.7, 1.0	0.1, 0.2, 0.3, 0.5, 0.7, 1.0			
LFH	, ,	, ,	, ,	, , ,	, ,			, , ,	
LFH Thickness	3	0 (forage)	0 (forage)	3	0 (forage)	0	5	0 (forage)	3
LFH pH (CaCl ₂)	6.3	-	-	4.3	-	-	5.9	-	6.0
LFH pH (H ₂ O)	6.6	-	-	4.7	-	-	6.2	-	6.3
LFH Bases (kmol _c ha ⁻¹)	22	-	-	9	-	-	39	-	17
LFH CEC (kmol _c ha ⁻¹)	23	-	-	22	-	-	43	-	19
Activity Coefficient of AI ¹⁺	0.872	-	-	0.923	-	-	0.896	-	0.902
Activity Coefficient of Al ²⁺	0.578	-	-	0.728	-	-	0.643	-	0.662
Activity Coefficient of Al ³⁺	0.291	-	-	0.490	-	-	0.371	-	0.395
Slope of pH-BSat Equation	3.76	3.76	3.76	3.76	3.76	3.76	3.76	3.76	3.76
Mineral 0-25 cm									
pH (CaCl ₂)	4.8	5.7	5.4	4.5	5.7	5.7	5.8	6.2	5.5
pH (H ₂ O)	5.2	6.1	5.8	5.0	6.1	6.1	6.2	6.6	5.9
Bases (kmol _c ha⁻¹)	127	309	106	92	618	1110	307	176	562
CEC (kmol _c ha ⁻¹)	238	376	164	202	739	1294	350	299	703
Activity Coefficient of Al	0.952	0.945	0.943	0.959	0.913	0.914	0.924	0.940	0.933
Activity Coefficient of Al ²⁺	0.822	0.797	0.791	0.847	0.695	0.697	0.727	0.790	0.757
Activity Coefficient of Al ³⁺	0.644	0.601	0.591	0.690	0.441	0.444	0.489	0.590	0.535
Slope of pH-BSat Eqn.	3.38	3.38	3.38	2.06	3.38	3.38	3.38	3.38	3.38
CO ₂ Partial Pressure (atm)	0.005	0.005	0.005	0.005	0.005	0.005	0.005	0.005	0.005
Weathering (kmol _c ha ⁻¹ yr ⁻¹)	0.10	0.10	0.07	0.07	0.15	0.10	0.10	0.10	0.15

Table 12. Input Data for Soil Acidification Simulations with the ARC Model (concluded)

5.3.2 Computations

The loss of bases is calculated on an annual basis from,

$$S = I - A - C - W \tag{17}$$

where S is the sum of bases lost, I is the effective acidity in the precipitation plus dryfall (the PAI), A is the acid leached out of the top 25 cm of soil, C is the decrease in bicarbonate weathering due to the decrease in soil solution pH, and W is the base contribution due to weathering. At the end of each year of simulation, a new sum of bases is calculated from the sum for the previous year. New values for pH, Al³⁺ concentration and BC:Al ratio are also calculated from equations relating pH with base saturation, pH with solution Al³⁺ concentration and pH with BC:Al ratio. A linear function describes the relationship between pH and base saturation percentage of the soil. The functions have been determined previously for mineral soil orders and reported by Abboud and Turchenek (1990) and for LFH layers by Abboud et al. (2002).

5.3.3 Changes to the ARC Model

Several changes were made to the earlier ARC model when applied in the Oil Sands area (Abboud et al, 2002) and to the Edmonton West grid Cell (Abboud and Turchenek, 2007). These included the addition of a new equation describing Al solubility in mineral soils and a new module to calculate the changes in mineral soil BC:Al ratios with changes in soil pH.

Al Solubility

The solubility of AI in the ARC model was assumed to follow the empirical model of Bloom and Grigal (1985), derived from Minnesota soils data. Recent changes to the ARC model, based on data from southeastern Alberta soils, resulted in the use of a more soluble form of gibbsite as an AI controlling mineral (Turchenek and Abboud, 2001). During our modeling of soil chemistry in the Oil Sands area, the solubility of AI in mineral horizons was further evaluated using archived data from a joint Syncrude-ARC project (Pauls et al 1996). The relationship between soluble AI and $pH(H_2O)$ derived from data in these projects was applied in the model to determine critical loads of soils.

Figure 1 shows the solubility relationship for mineral soils in the upper 25 cm in the Oil Sands region. These covered several soil orders in the area. A linear relationship is evident with a significant R^2 term. This equation is similar in form to the Bloom and Grigal (1985) and Turchenek and Abboud (2001) equations and seems to imply a strong role for a mineral form controlling Al solubility. The pH coefficient in the equation (2.66) is close to the theoretical 3 required for gibbsite to be a controlling mineral, and the constant term (8.10) is close to the theoretical 8 assumed for the solubility product of gibbsite. Thus the possibility of gibbsite controlling Al solubility in these soils is strong with the likelihood of some influence from the organic matter present in the Ah horizons and/or leaching from the LFH layer.

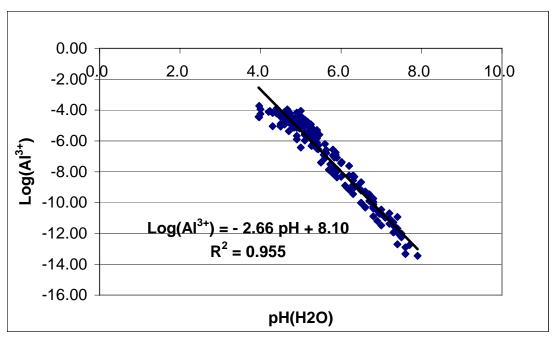


Figure 1. Al Solubility in Mineral Horizons

The pH-AI solubility relationship was similarly derived for the LFH layers of soils. In summary, the equations applied in modelling the soils of the Edmonton East grid cell were:

for LFH horizons:	$pH(H_2O) = -2.72 pH(H_2O) + 8.03$	R ² = 0.923, n= 65 samples	(14)
for 0-25 cm layer:	$pH(H_2O) = 2.66 pH(H_2O) + 8.10$	R ² = 0.955, n= 130 samples	(15)

BC:Al Ratios

The relationship between BC:AI ratios and pH for mineral soil layers was also derived from examination of soils in the oil sands region, as described in Abboud et al. (2002). An exponential relationship between BC:AI ratios and pH was observed as shown in Figure 2. This equation shows scatter that is likely due to the diverse nature of the soil orders and their mineralogy and texture, and to the influence of weathering and exchange/adsorption processes to both organic and mineral surfaces.

The pH-AI solubility relationship was derived for both the mineral and the LFH layers of soils. The equations applied in modelling the soils of the Edmonton East grid cell were:

for LFH horizons:	BC:AI Ratio = 0.12e ^{1.40pH(H2O)}	$R^2 = 0.576$, n= 65 samples	(16)
for 0-25 cm layer:	BC:AI Ratio = 0.043e ^{1.14pH(H2O)}	$R^2 = 0.641$, n= 65 samples	(17)

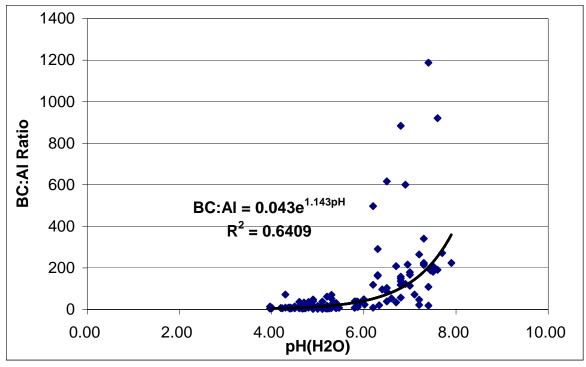


Figure 2. BC:Al Ratio in Mineral Horizons

5.3.4 Model Execution and Data Outputs

Computations were made for changes in soil properties on an annual basis. Output data for each time interval included: (1) year; (2) pH of soil; (3) acid input; (4) acid output; (5) protonation; (6) change in pH; (7) base saturation; (8) sum of base cations; (9) base cations lost; (10) Al³⁺ concentration in soil solution, and (11) BC:Al ratio.

The outputs of major interest are the changing values of soil pH, base saturation, and BC:Al during the time period selected. Model data were transferred to EXCEL spreadsheets to facilitate data analysis in terms of critical loads. Simulations were conducted with a desktop computer using the program RS1. Table 13 shows the model output information generated in a table.

5.3.5 Model Output

The ARC model predictions of critical loads for critical chemical values reached after 50 and 100 years of acid deposition were derived from tabulated model output data for a 300 year period, and are presented in Table 14. Table 14 shows the changes in relation to given acid deposition inputs for four time periods. An example of model outputs is also presented in diagrammatic form in the charts in Figure 3.

	Mineral Soil Layer														
Time	рН	Sol. Al	Base	BC:AI	Acid In	Acid Out	Weathering	Protonat.	Bases Lost	Exch. Bases	Soil				
(Years)	(H2O)	(M)	Saturation	Ratio		(kmol ha ⁻¹ yr ⁻¹)									
0	5.5	5.38E-07	0.73	60	1.0	0.0	0.00E+00	0.00E+00	0.0	370.4	Edward Plain 1				
1	5.5	5.38E-07	0.73	60	1.0	0.0	0.00E+00	0.00E+00	1.0	369.4	Edward Plain 1				
2	5.5	5.27E-07	0.73	60	1.0	0.0	1.30E-04	4.98E-04	1.0	368.4	Edward Plain 1				
3	5.5	5.16E-07	0.72	59	1.0	0.0	2.60E-04	9.91E-04	1.0	367.4	Edward Plain 1				

Table 13. Example of Output from the ARC Model Simulating Processes

Table 14. Changes in Soil Chemistry in Relation to Different Acid Inputs

Acid Input		Mineral pH _h				eral Bas	e Satura	tion	Mineral BC:Al Ratio			
kmol ha ⁻¹ yr ⁻¹	0 yr	50 yr	100 yr	300 yr	0 yr	50 yr	100 yr	300 yr	0 yr	50 yr	100 yr	300 yr
						Islet U	pland 1					
0.1	5.1	5.1	5.1	5.1	0.70	0.70	0.70	0.70	15	15	15	15
0.2	5.1	5.1	5.1	5.1	0.70	0.70	0.70	0.70	15	15	15	15
0.3	5.1	5.1	5.1	5.1	0.70	0.70	0.70	0.70	15	15	15	15
0.5	5.1	5.1	5.1	5.1	0.70	0.70	0.70	0.70	15	15	15	15
0.7	5.1	5.1	5.1	5.0	0.70	0.70	0.70	0.64	15	15	15	13
1	5.1	5.1	5.1	4.7	0.70	0.70	0.67	0.64	15	15	14	9
						Islet U	pland 2					-
0.1	4.6	4.6	4.6	4.6	0.34	0.34	0.34	0.34	8	8	8	8
0.2	4.6	4.6	4.6	4.6	0.34	0.34	0.34	0.34	8	8	8	8
0.3	4.6	4.6	4.6	4.6	0.34	0.34	0.34	0.34	8	8	8	8
0.5	4.6	4.6	4.6	4.6	0.34	0.34	0.34	0.34	8	8	8	8
0.7	4.6	4.6	4.6	4.6	0.34	0.34	0.34	0.34	8	8	8	8
1	4.6	4.6	4.5	4.5	0.34	0.34	0.32	0.31	8	8	8	7
	Islet Upland 3								-			
0.1	6.9	6.9	6.9	6.9	0.98	0.98	0.98	0.97	115	114	113	112
0.2	6.9	6.9	6.9	6.9	0.98	0.97	0.98	0.96	115	112	111	109
0.3	6.9	6.9	6.9	6.8	0.98	0.96	0.95	0.94	115	107	107	103
0.5	6.9	6.9	6.8	6.6	0.98	0.94	0.91	0.86	115	103	95	84
0.7	6.9	6.8	6.7	6.5	0.98	0.92	0.88	0.80	115	93	88	71
1	6.9	6.7	6.6	6.2	0.98	0.89	0.83	0.68	115	91	78	53
						Islet U	pland 4					-
0.1	5.5	5.5	5.5	5.5	0.83	0.83	0.83	0.83	23	23	23	23
0.2	5.5	5.5	5.5	5.5	0.83	0.83	0.83	0.83	23	23	23	23
0.3	5.5	5.5	5.5	5.5	0.83	0.83	0.83	0.83	23	23	23	23
0.5	5.5	5.5	5.5	5.5	0.83	0.83	0.83	0.83	23	23	23	23
0.7	5.5	5.5	5.5	5.5	0.83	0.83	0.83	0.83	23	23	23	23
1	5.5	5.5	5.5	5.5	0.83	0.83	0.83	0.83	23	23	23	23
						Islet U	pland 5					-
0.1	5.6	5.6	5.5	5.2	0.64	0.64	0.61	0.52	26	26	23	16
0.2	5.6	5.6	5.5	5.2	0.64	0.64	0.61	0.52	26	26	23	16
0.3	5.6	5.6	5.5	5.2	0.64	0.64	0.61	0.52	26	26	23	16
0.5	5.6	5.6	5.5	5.2	0.64	0.64	0.61	0.52	26	26	23	16
0.7	5.6	5.6	5.5	5.2	0.64	0.64	0.61	0.52	26	26	23	16
1	5.6	5.6	5.5	5.2	0.64	0.64	0.61	0.52	26	26	23	16

Acid Input		Miner	al pH _h		Min	eral Bas	e Satura	tion	N	lineral B	C:Al Rati	Al Ratio	
kmol ha ⁻¹ yr ⁻¹	0 yr	50 yr	100 yr	300 yr	0 yr	50 yr	100 yr	300 yr	0 yr	50 yr	100 yr	300 yr	
		1			Eldo	rena/Re	dwater P	lains		1	1	1	
0.1	5.0	4.9	4.9	4.8	0.47	0.42	0.39	0.37	14	12	11	11	
0.2	5.0	4.9	4.9	4.8	0.47	0.42	0.39	0.37	14	12	11	11	
0.3	5.0	4.9	4.8	4.8	0.47	0.41	0.38	0.36	14	12	11	10	
0.5	5.0	4.8	4.7	4.7	0.47	0.35	0.31	0.30	14	10	9	9	
0.7	5.0	4.7	4.6	4.6	0.47	0.30	0.27	0.27	14	9	8	8	
1	5.0	4.6	4.5	4.5	0.47	0.24	0.23	0.23	14	8	8	8	
		1	1			Edward	Plain 1			1	1	1	
0.1	5.5	5.5	5.5	5.5	0.73	0.73	0.73	0.72	23	23	23	23	
0.2	5.5	5.5	5.5	5.4	0.73	0.73	0.72	0.69	23	23	23	21	
0.3	5.5	5.5	5.5	5.3	0.73	0.73	0.71	0.64	23	23	22	19	
0.5	5.5	5.5	5.4	5.1	0.73	0.72	0.68	0.55	23	22	21	15	
0.7	5.5	5.4	5.3	4.9	0.73	0.70	0.65	0.45	23	22	19	12	
1	5.5	5.4	5.2	4.7	0.73	0.68	0.59	0.33	23	20	17	9	
	Edward Plain 2										1		
0.1	5.4	5.4	5.4	5.3	0.63	0.63	0.62	0.59	21	20	20	19	
0.2	5.4	5.4	5.3	5.2	0.63	0.62	0.60	0.52	21	20	19	16	
0.3	5.4	5.4	5.3	5.0	0.63	0.61	0.57	0.45	21	20	18	13	
0.5	5.4	5.3	5.1	4.8	0.63	0.58	0.51	0.33	21	18	15	10	
0.7	5.4	5.2	5.0	4.6	0.63	0.55	0.44	0.26	21	17	13	9	
1	5.4	5.1	4.8	4.6	0.63	0.50	0.35	0.22	21	15	11	8	
		1				Watt La	ke Plain	1		1	1	1	
0.1	6.3	6.3	6.3	6.2	0.98	0.97	0.97	0.95	57	55	54	51	
0.2	6.3	6.3	6.2	6.1	0.98	0.96	0.95	0.92	57	54	52	46	
0.3	6.3	6.2	6.2	6.0	0.98	0.96	0.94	0.89	57	53	49	41	
0.5	6.3	6.2	6.1	5.8	0.98	0.94	0.92	0.82	57	50	45	32	
0.7	6.3	6.1	6.0	5.6	0.98	0.93	0.89	0.76	57	48	41	24	
1	6.3	6.1	5.9	5.2	0.98	0.91	0.85	0.65	57	44	35	16	
			•		Ν	lusidora	Upland	1		1	1	•	
0.1	5.2	5.2	5.2	5.2	0.53	0.53	0.53	0.53	16	16	16	16	
0.2	5.2	5.2	5.2	5.2	0.53	0.53	0.53	0.53	16	16	16	16	
0.3	5.2	5.2	5.2	5.2	0.53	0.53	0.53	0.53	16	16	16	16	
0.5	5.2	5.2	5.2	5.2	0.53	0.53	0.53	0.53	16	16	16	16	
0.7	5.2	5.2	5.2	5.2	0.53	0.53	0.53	0.53	16	16	16	16	
1	5.2	5.2	5.2	5.2	0.53	0.53	0.53	0.53	16	16	16	16	
		1	1		Ν	/lusidora	Upland 2	2				1	
0.1	6.1	6.1	6.0	6.0	0.82	0.81	0.80	0.78	45	43	42	39	
0.2	6.1	6.0	6.0	5.8	0.82	0.80	0.78	0.73	45	41	39	32	
0.3	6.1	6.0	5.9	5.6	0.82	0.79	0.76	0.68	45	40	36	26	
0.5	6.1	5.9	5.7	5.2	0.82	0.76	0.72	0.56	45	36	30	17	
0.7	6.1	5.8	5.6	4.8	0.82	0.74	0.67	0.44	45	33	25	10	
1	6.1	5.7	5.3	4.2	0.82	0.70	0.59	0.25	45	29	19	5	

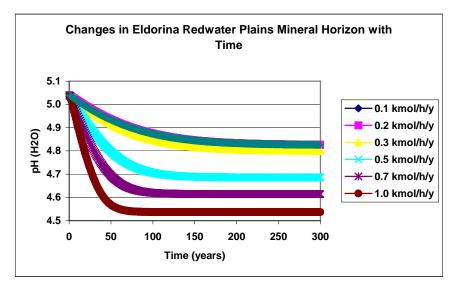
Table 14. Changes in Soil Chemistry in Relation to Different Acid Inputs

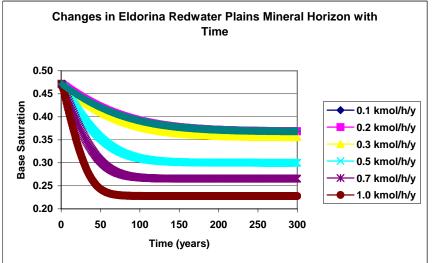
Acid Input		Miner	al pH _h		Min	eral Bas	e Satura	tion	N	lineral B	C:Al Rati	0
kmol ha⁻¹ yr⁻¹	0 yr	50 yr	100 yr	300 yr	0 yr	50 yr	100 yr	300 yr	0 yr	50 yr	100 yr	300 yr
		1	I		Ν	lusidora	Upland	3		1		
0.1	5.8	5.7	5.7	5.5	0.65	0.63	0.61	0.57	32	30	28	24
0.2	5.8	5.6	5.5	5.1	0.65	0.60	0.56	0.45	32	27	23	15
0.3	5.8	5.5	5.3	4.7	0.65	0.57	0.51	0.33	32	24	19	9
0.5	5.8	5.4	5.0	4.0	0.65	0.52	0.40	0.13	32	19	13	4
0.7	5.8	5.2	4.6	3.7	0.65	0.46	0.30	0.03	32	16	8	3
1	5.8	4.9	4.1	3.6	0.65	0.38	0.15	0.00	32	11	5	3
					Ν	lusidora	Upland	4				
0.1	4.9	4.8	4.8	4.8	0.45	0.42	0.40	0.38	12	11	10	10
0.2	4.9	4.8	4.8	4.8	0.45	0.42	0.40	0.38	12	11	10	10
0.3	4.9	4.8	4.8	4.8	0.45	0.42	0.40	0.38	12	11	10	10
0.5	4.9	4.8	4.7	4.7	0.45	0.40	0.37	0.35	12	10	9	9
0.7	4.9	4.7	4.6	4.6	0.45	0.36	0.33	0.32	12	9	9	8
1	4.9	4.6	4.5	4.5	0.45	0.32	0.28	0.28	12	8	8	8
	Musidora Upland 5											
0.1	6.1	6.1	6.1	6.0	0.84	0.83	0.83	0.81	45	44	43	41
0.2	6.1	6.1	6.0	5.9	0.84	0.82	0.81	0.78	45	43	41	36
0.3	6.1	6.0	6.0	5.8	0.84	0.82	0.80	0.75	45	42	39	32
0.5	6.1	6.0	5.9	5.6	0.84	0.81	0.78	0.68	45	40	36	24
0.7	6.1	5.9	5.8	5.3	0.84	0.79	0.75	0.61	45	38	32	19
1	6.1	5.9	5.7	5.0	0.84	0.77	0.71	0.50	45	35	28	12
		1	I			Whitfo	rd Plain			I	I	I
0.1	6.1	6.1	6.1	6.0	0.86	0.85	0.85	0.84	45	44	44	42
0.2	6.1	6.1	6.1	6.0	0.86	0.86	0.84	0.82	45	44	43	39
0.3	6.1	6.1	6.0	5.9	0.86	0.85	0.84	0.80	45	43	42	36
0.5	6.1	6.0	6.0	5.8	0.86	0.84	0.82	0.76	45	42	39	31
0.7	6.1	6.0	5.9	5.6	0.86	0.83	0.81	0.72	45	41	37	26
1	6.1	6.0	5.9	5.4	0.86	0.82	0.79	0.65	45	39	34	20
		1	1			Delph	Upland			1	1	1
0.1	6.2	6.2	6.2	6.2	0.88	0.88	0.88	0.88	52	52	52	52
0.2	6.2	6.2	6.2	6.2	0.88	0.88	0.88	0.88	52	52	51	51
0.3	6.2	6.2	6.2	6.2	0.88	0.88	0.88	0.87	52	51	51	51
0.5	6.2	6.2	6.2	6.0	0.88	0.88	0.87	0.83	52	51	50	42
0.7	6.2	6.2	6.0	5.7	0.88	0.87	0.83	0.72	52	50	43	28
1	6.2	6.1	5.8	5.0	0.88	0.84	0.77	0.54	52	45	33	14
			<u> </u>			Norma	Plain 1					
0.1	6.5	6.5	6.4	6.4	0.59	0.58	0.57	0.57	71	68	67	65
0.2	6.5	6.4	6.4	6.3	0.59	0.56	0.55	0.54	71	65	62	59
0.3	6.5	6.4	6.3	6.2	0.59	0.55	0.53	0.50	71	62	57	52
0.5	6.5	6.3	6.2	5.9	0.59	0.53	0.49	0.42	71	56	49	37
0.7	6.5	6.2	6.0	5.6	0.59	0.50	0.44	0.31	71	51	41	24

Table 14. Changes in Soil Chemistry in Relation to Different Acid Inputs

Acid Input		Miner	al pH _h		Min	eral Bas	e Satura	tion	N	lineral B	C:Al Rati	o	
kmol ha ⁻¹ yr ⁻¹	0 yr	50 yr	100 yr	300 yr	0 yr	50 yr	100 yr	300 yr	0 yr	50 yr	100 yr	300 yr	
1	6.5	6.1	5.8	4.9	0.59	0.46	0.37	0.11	71	44	31	11	
		Norma Plain 2											
0.1	5.9	5.9	5.9	5.9	0.80	0.80	0.80	0.80	37	37	37	37	
0.2	5.9	5.9	5.9	5.9	0.80	0.80	0.80	0.80	37	37	37	37	
0.3	5.9	5.9	5.9	5.9	0.80	0.80	0.80	0.80	37	37	37	37	
0.5	5.9	5.9	5.9	5.9	0.80	0.80	0.80	0.79	37	37	36	36	
0.7	5.9	5.9	5.8	5.7	0.80	0.79	0.78	0.75	37	36	35	30	
1	5.9	5.8	5.7	5.4	0.80	0.78	0.75	0.65	37	34	30	20	

Table 14. Changes in Soil Chemistry in Relation to Different Acid Inputs





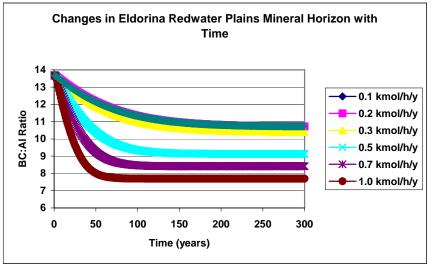


Figure 3. Example of Soil Chemistry Changes Over Time

5.3.6 Critical Chemical Values for the Soil Groups

The rational for selecting critical chemical values was described in Section 2.2. In order to derive critical loads, the threshold level for a change in a chemical parameter must be selected, as well as the period over which this change can occur. Once a threshold is reached, however, soil chemistry would be negatively affected. A level of protection of the soil was considered whereby only a percentage of a parameter in question would be affected, well before the critical load is reached. This approach was applied in developing critical loads in the Oil Sands area, upon the suggestion of the NOx/SO2 Management Working Group of the Cumulative Environmental Management Association (Abboud et al., 2002). Levels of 75% of the original soil value, and the mid-point between the original value and the literature-based critical load, were examined. The more protective of these levels is recommended in the examination of critical loads for the Edmonton East grid cell, namely application of the 75% case as the soil critical load. Acidification (i.e., PAI) levels resulting in these thresholds being reached within 50 and 100 years were derived from the model data. The lower value of the BC:AI ratio or base saturation percentage is suggested as the critical load. PAI levels required to reach critical chemical values were also examined, including the critical values of $pH(H_2O)$ 4.0 and 5.6, for forested and grassland soils respectively. Table 15 shows the critical chemical values established or calculated for the various soil groups.

5.3.7 Critical Load Derivation

The time frame within which changes in soil chemistry occur is an important consideration in using dynamic models to derive critical loads. Decisions are required as to whether critical chemical values of soil chemical parameters may be reached in only a few years, or over a longer period. Fifty and one hundred year time periods were selected for these decisions. Fifty years is a relatively short period, and its selection is based on the view that it is of sufficient length to enable detection of an actual acidification trend and to initiate measures to counteract the trend. One hundred years is a longer time frame that results in a lower critical load, and it therefore provides a greater measure of protection.

The ARC model predictions of critical loads for critical chemical values reached after 50 and 100 years of acid deposition were derived from the tabulated model output (Table 14), and are presented in Table 16.

The lowest critical loads were obtained for the 75% case of the BC:AI ratio. Applying the principle of selecting the lowest of the calculated critical loads, the BC:AI ratio would therefore form the basis of critical loads for soils in the study area. With some exceptions, the highest critical loads were obtained for the Luvisolic and Chernozemic soils. Although coarse textured Chernozemic soils were selected in this investigation, the relatively high organic mater content provides a large supply of exchangeable cations in addition to that associated with the mineral component alone. Luvisols are likely influenced by finer textures as compared to the Brunisols and Chernozems. The Brunisols in the Edmonton East area have low organic matter and clay contents, and therefore the least acid buffering capacity.

		-	pHh ₌	Mineral (0-25 cm) Critical Chemical Value						
Site	Soil Series	pHh _i ^z	4.0 or 5.6	BSat _i	BSat _i x 0.75	BSat =0.1, 0.5	BC:Al _i	BC:Al _i x 0.75	BC:AI =1, 45	
Luvisols										
Islet Upland 1	Cooking Lake	5.1	4.0	0.70	0.53	0.1	15	11	1	
Islet Upland 2	Cooking Lake-AC	4.6	4.0	0.34	0.26	0.1	8	6	1	
Islet Upland 3	Uncas (Forage)	6.9	4.0	0.98	0.74	0.1	115	86	1	
Islet Upland 4	Uncas	5.5	4.0	0.75	0.56	0.1	23	17	1	
Brunisols										
Eldorena/Redwater Plains	Nestow	5.0	4.0	0.47	0.35	0.1	14	11	1	
Edward Plain 1	Nestow	5.5	4.0	0.73	0.55	0.1	23	17	1	
Edward Plain 2	Nestow	5.4	4.0	0.63	0.47	0.1	21	16	1	
Musidora Upland 4	Nestow	5.0	4.0	0.45	0.34	0.1	12	9	1	
Chernozems										
Islet Upland 5	Helliwell-GR	5.6	5.6	0.64	0.48	0.5	26	20	45	
Musidora Upland 1	Helliwell	5.2	5.6	0.53	0.40	0.5	16	12	45	
Musidora Upland 2	Helliwell (Forage)	6.1	5.6	0.82	0.62	0.5	45	34	45	
Musidora Upland 3	Helliwell (Forage)	5.8	5.6	0.65	0.49	0.5	32	24	45	
Musidora Upland 5	Peace Hills	6.1	5.6	0.84	0.63	0.5	45	34	45	
Watt Lake Plain	Mundare (Forage)	6.3	5.6	0.98	0.74	0.5	57	43	45	
Whitford Plain	Ukalta (Native)	6.1	5.6	0.86	0.65	0.5	45	34	45	
Delph Upland	Helliwell	6.2	5.6	0.88	0.66	0.5	52	39	45	
Norma Plain 1	Mundare (Forage)	6.5	5.6	0.59	0.44	0.5	71	53	45	
Norma Plain 2	Ukalta	5.9	5.6	0.80	0.60	0.5	37	28	45	

Table 15. Critical Chemical Values Calculated from Initial Soil Data

² Abbreviations: pH_h - soil pH measured in H₂O; pH_{hi} - initial pH_h; BSat - base saturation percentage; BSat_i - initial BSat; BC:AI - base cation to aluminum ratio in soil solution; BC:AI_i - initial BC:AI; GR - gravelly; AC - acidic

Site	Soil Series	Time	Mineral Critical Load Value ^z (kmol H⁺ ha ⁻¹ yr ⁻¹)						
One	oon oenes	(years)	рН _h 4.0, 5.6	BSat _i x 0.75	BSat= 0.1, 0.5 ^Y	BC:Al _i x 0.75	BC:Al= 1 or 45		
Luvisols									
Islet Upland 1	Cooking Lake	50	>1	>1	>1	>1	>1		
		100	>1	>1	>1	>1	>1		
Islet Upland 2	Cooking Lake (Acidic)	50	>1	>1	>1	>1	>1		
		100	>1	>1	>1	>1	>1		
Islet Upland 3	Uncas (Forage)	50	>1	>1	>1	>1	>1		
		100	>1	>1	>1	0.6	>1		
Islet Upland 4	Uncas	50	>1	>1	>1	>1	>1		
		100	>1	1	>1	>1	>1		
Brunisols									
Eldorena/ Redwater Plains	Nestow	50	>1	0.5	>1	0.4	>1		
		100	>1	0.4	>1	0.3	>1		
Edward Plain 1	Nestow	50	>1	>1	>1	>1	>1		
		100	>1	0.5	>1	1	>1		
Edward Plain 2	Nestow	50	>1	>1	>1	0.9	>1		
		100	>1	0.6	>1	0.4	>1		
Musidora Upland 4	Nestow	50	>1	0.8	>1	0.7	>1		
		100	>1	0.6	>1	0.5	>1		
Chernozems									
Islet Upland 5	Nestow (gravelly)	50	>1	>1	>1	>1	<1 ^X		
		100	>1	>1	>1	>1	<1		
Musidora Upland 1	Nestow	50	<1	>1	>1	>1	<1		
		100	<1	>1	>1	>1	<1		
Musidora Upland 2	Helliwell (Forage)	50	>1	>1	>1	0.6	<1		
		100	0.7	0.8	>1	0.4	<1		
Musidora Upland 3	Nestow (Forage)	50	0.2	0.6	0.6	0.15	<1		
		100	0.15	0.3	0.3	0.1	<1		
Musidora Upland 5	Peace Hills	50	>1	>1	>1	>1	<1		
		100	>1	>1	>1	0.6	<1		
Watt Lake Plain	Mundare (Forage)	50	>1	>1	>1	>1	0.9		
		100	>1	>1	>1	0.6	0.5		
Whitford Plain	Ukalta (Native)	50	>1	>1	>1	>1	<1		
		100	>1	>1	>1	1	<1		
Delph Upland	Helliwell	50	>1	>1	>1	>1	1		
		100	>1	>1	>1	0.8	0.6		
Norma Plain 1	Mundare (Forage)	50	>1	>1	0.7	0.6	1		
	、 、 ,	100	>1	0.7	0.5	0.4	0.6		
Norma Plain 2	Ukalta	50	>1	>1	>1	>1	<1		
		100	>1	>1	>1	>1	<1		

Table 16. ARC Model Predictions of Critical Loads for Critical Chemical Values Reached after 50 and 100 Years of Acid Deposition

pH_h critical values are 4.0 for Luvisols and Brunisols, and 5.6 for Chernozems; BC:Al critical values (last column) are 1 for Luvisols and Brunisols, and 45 for Chernozems.
 BSat=0.5 applied to Chernozemic soils

z <1 indicated where the initial BC:Al ratio (BC:Ali) is already below the threshold value of 45 for Chernozemic soils.

5.4 COMPARISON OF METHODS OF CRITICAL LOAD DERIVATION

Critical loads for sandy soils in the Edmonton East study area, as determined by the empirical Skokloster approach, Steady State Mass Balance, and ARC models are summarized in Table 17.

Critical Load					Critical	Loa	d (kmo	IH⁺ ha⁻¹	yr ⁻¹))			
Derivation	Luvisol Land Systems						Brunisol Land Systems						
Method and Criterion	Islet Upland 1	Islet Upland	Islet 2 Upland		Isle [.] Uplan		Red	orena/ water ains	_	Edward Plain 1	Edwar Plain		usidora pland 4
Sites	13, 16, 18	3 17	11		12		5, 6, 7	7, 8, BR		3	4	9, 1	0, 23, 26
Skokloster	1-2	1-2	1-2		1-2		0.2	2-0.5		0.5-1	0.2-0.	5 ().2-0.5
SSMB	2.9	2.9	2.9		2.9		C).6		0.7	0.6		0.6
ARC pH₅₀	>1	>1	>1		>1			>1		>1	>1		>1
ARC pH ₁₀₀	>1	>1	>1		>1			>1		>1	>1		>1
ARC BSat ₅₀	>1	>1	>1		>1		C).5		>1	>1		0.8
ARC BSat ₁₀₀	>1	>1	>1		1		C).4		0.5	0.6		0.6
ARC BC:AI ₅₀	>1	>1	>1		>1		C).4		>1	0.9		0.7
ARC BC:AI100	>1	>1	0.6		>1		C).3		1	0.4		0.5
Critical Load					Critical	Load	d (kmo	l H⁺ ha⁻¹	yr ⁻¹))			
Derivation					Cher	noze	m Lane	d Systen	ns				
Method and	Islet	Musidora	Musidora	Mu	sidora	Mus	sidora	Watt La	ke	Whitford	Delph	Norma	Norma
Criterion	Upland 5	Upland 1	Upland 2	Up	land 3	Upl	and 5	Plain		Plain	Upland	Plain 1	Plain 2
Sites	19	24	22		25		15	1		2	21	TH1	20
Skokloster	0.5-1	0.2-0.5	0.2-0.5	0.	2-0.5	0.	.5-1	0.5-1		0.5-1	0.5-1	0.2-0.5	0.5-1
SSMB	0.7	0.7	0.7		0.7	1	1.0	1.0		1.0	1.0	0.7	1.0
ARC pH₅₀ ^z	>1	-	>1		0.2	2	>1	>1		>1	>1	>1	>1
ARC pH ₁₀₀ ²	>1	-	0.7	(0.15	;	>1	>1		>1	>1	>1	>1
ARC BSat ₅₀	>1	>1	>1		0.6	;	>1	>1		>1	>1	>1	>1
ARC BSat ₁₀₀	>1	>1	0.8		0.3	;	>1	>1		>1	>1	0.7	>1
ARC BC:Al ₅₀	>1	>1	0.6	(0.15	;	>1	>1		>1	>1	0.6	>1
ARC BC:Al ₁₀₀	>1	>1	0.4		0.1	().6	0.6		1	0.8	0.4	>1

Table 17. Comparison of Critical Loads Derived by Different Methods

² Subscripts 50 and 100 – load required to reach a critical chemical threshold after 50 and 100 years, respectively; e.g., for the Musidora Upland 2, the critical load is >1 kmol H⁺ ha⁻¹ yr⁻¹ to reach pH in 50 years, but 0.7 kmol H⁺ ha⁻¹ yr⁻¹ to reach pH 5.6 in 100 years.

The highest critical loads were obtained with the Skokloster and the SSMB approaches for the Islet Upland 1, 2 3 and 4 soils. These soils are all Luvisols with clay loam parent materials. These approaches account for the weathering capacity of these soils, with the SSMB model considering weathering to a depth of 75 cm. Lower critical loads were obtained with the ARC model which considers only the top 25 cm of the soil. For Luvisols and Brunisols, there is generally good agreement among the three methods, with particularly good agreement between the ARC 100 year critical loads and the SSMB critical loads.

Critical loads for Chernozemic soils based on the ARC model tend be higher than those determined by the Skokloster and SSMB methods, although not in all cases. There is additional acid buffering capacity in the Chernozemic soils due to exchangeable cations associated with relatively high organic matter contents. This factor is not accounted for in the Skokloster and SSMB methods.

The lowest critical loads were obtained for the ARC BC:AL₁₀₀ criterion. Critical loads according to the ARC BC:AL₅₀ criterion were commonly similar to those of the Skokloster approach, while the SSMB CLs were consistently higher than both of these. However, the SSMB CLs were generally lower than those based on the ARC pH criteria.

Both the Skokloster and SSMB approaches to setting critical loads are based on maintaining steady-state over a very long time. They are based on replenishment of base cations in soil by weathering. A considerable amount of buffering capability is provided by cations on the cation exchange complex. For protection of soils in the relatively short term, it would be appropriate to simulate soil chemistry based on both cation exchanges buffering and weathering than on methods based on weathering alone. The ARC model applies these processes and additionally enables the examination of changes in soil chemistry over time.

6.0 ACIDIFICATION SENSITIVITY

6.1 SENSITIVITY CLASSES

Previous sections of this report have focused on deriving the critical load for individual soil profiles or groups of very similar profiles. The profiles for which critical loads were derived in this study can be considered to be representative of the various soil series examined. For mapping purposes, however, the critical loads were considered in terms of sensitivity classes by applying an approach developed for the Provost-Esther critical loads study (Turchenek and Abboud, 2001). The approach uses both 50 and 100 year model results, and links the critical load determinations to sensitivity classes and to mapping of the loads.

The critical loads were assigned to a sensitivity class that could more or less be equated with critical loads for application in Alberta (Clean Air Strategic Alliance and Alberta Environment 1999). These critical loads are 0.25 kmol ha⁻¹ yr⁻¹ for sensitive soils, 0.50 kmol ha⁻¹ yr⁻¹ for moderately sensitive soils, and 1.00 kmol ha⁻¹ yr⁻¹ for low sensitivity soils. Turchenek and Abboud (2001) suggested critical load and sensitivity classes as follows:

\leq 0.2 kmol ha ⁻¹ yr ⁻¹ ; critical chemical value reached within 100 years 0.2 to 0.5 kmol ha ⁻¹ yr ⁻¹ ; critical chemical value within 50 years	Sensitive Sensitive
0.2 to 0.5 kmol ha ⁻¹ yr ⁻¹ ; critical chemical value within 100 years	Moderate sensitivity
0.5 to 1.0 kmol ha ⁻¹ yr ⁻¹ ; critical chemical value within 50 years 0.5 to 1.0 kmol ha ⁻¹ yr ⁻¹ ; critical chemical value within 100 years	Moderate sensitivity Low sensitivity
>1.0 kmol ha ⁻¹ yr ⁻¹ ; critical chemical value within 50 years	Low sensitivity

The more stringent of the base saturation or BC:Al critical loads obtained by modelling was used to determine the sensitivity category. The above categories of soil sensitivity indicate, for example, that if BSat or BC:Al is reduced to 75% of the original value within 100 years at a Potential Acid Input level of ≤ 0.2 kmol ha⁻¹ yr⁻¹, than the soil would be regarded as Sensitive. If 0.2 to 0.5 kmol ha⁻¹ yr⁻¹ reduces these soil properties to the critical chemical values within 50 years, then the soil would also be regarded as Sensitive. However, if 50 to 100 years is required at this latter level, than the soil would be allocated to the Moderate sensitivity class.

This approach enables the allocation of a specific soil profile to a sensitivity class. The above criteria were applied to the eighteen representative soils or soil groups to which the ARC model was applied, and compared to acidification sensitivity criteria of Holowaychuk and Fessenden (1987). Results are presented in Table 18, along with Holowaychuk and Fessenden ratings.

Land System	Soil Series	pHc ^z	рН _h	Texture 0-25 cm	50 Yr CL	100 Yr CL	H-F Class	Sensitivity Class
Luvisols								
Islet Upland 1	Cooking Lake	4.7	5.1	Sandy Loam	>1	>1	М	L
Islet Upland 2	Cooking Lake (Acidic)	4.1	4.6	Sandy Loam	>1	>1	М	L
Islet Upland 3	Uncas (Forage)	6.6	6.9	Sandy Loam	>1	0.6	М	L
Islet Upland 4	Uncas	5.1	5.5	Sandy Loam	>1	>1	М	L
Brunisols								
Eldorena/ Redwater Plains	Nestow	4.6	5.0	Sand	0.4	0.3	L	S
Edward Plain 1	Nestow	5.1	5.5	Loamy Sand	>1	1	L	L
Edward Plain 2	Nestow	5.0	5.4	Sand	0.9	0.4	L	М
Musidora Upland 4	Nestow	4.5	5.0	Sand	0.7	0.5	L	М
Chernozems								
Islet Upland 5	Helliwell (gravelly)	5.2	5.6	Loamy Sand	>1	>1	L	L
Musidora Upland 1	Helliwell (Native)	4.8	5.2	Loamy Sand	>1	>1	L	L
Musidora Upland 2	Helliwell (Forage)	5.7	6.1	LS-SL	0.6	0.4	L	М
Musidora Upland 3	Helliwell (Forage)	5.4	5.8	Loamy Sand	0.15	0.1	L	S
Musidora Upland 5	Peace Hills	5.7	6.1	Loamy Sand	>1	0.6	L	L
Watt Lake Plain	Mundare (Forage)	5.9	6.3	Loamy Sand	>1	0.6	L	L
Whitford Plain	Ukalta (Native)	5.7	6.1	Loamy Sand	>1	1	L	L
Delph Upland	Helliwell	5.8	6.2	Sandy Loam	>1	0.8	L	L
Norma Plain 1	Mundare (Forage)	6.1	6.5	Sand	0.6	0.4	L	М
Norma Plain 2	Ukalta	5.5	5.9	Sandy Loam	>1	>1	L	L

 Table 18. Critical Loads and Derived Sensitivity Classes

² Abbreviations: pH_c - pH(CaCl₂); pH_h - pH(H₂O); CL - critical load; H-F Class - Holowaychuk-Fessenden sensitivity class

The H-F sensitivity rating in the above table is based directly on the sensitivity map of Holowaychuk and Fessenden (1987). On this map, only the Cooking Lake moraine (i.e., the Islet Upland Land System) was identified as having soils with potentially medium sensitivity. The ARC modelling results suggest that the predominantly Luvisolic soils of the Islet Upland have

Low sensitivity to acidification. Other differences between the Holowaychuk-Fessenden mapping and the ARC model results pertain to the sandy Brunisolic soils of areas such as Redwater Plain, Eldorena Plain and Musidora Upland. These are mapped as being of Low Sensitivity in the Holowaychuk-Fessenden map. In the current study, some of the soils that characterize these Land Systems were indicated as having High and Medium sensitivity according to the ARC model.

Differences in the sensitivity classes are likely due to a variety of factors. In the case of the sandy soil areas, the H-F map legend indicates that the soils have high organic matter content and high exchangeable cation content. These properties are associated with the finer textured soils in these Land Systems. However, there appears to have been very limited soil profile information for the sandiest soils in the region. These were mapped as 'Dune Sand' soils in the soil survey of the Edmonton sheet (Bowser et al. 1962).

Since the publication of the acidification sensitivity map by Holowaychuk and Fessenden (1987), more detailed mapping has been carried out for the AGRASID (Agricultural Region of Alberta Soil Information Database) program (Alberta Soil Information Centre, 2007). The sandy soils on eolian deposits were mapped as Eluviated Eutric Brunisols, Primula Soil Series. In the current study, the soil profiles sampled were Eluviated Dystric Brunisols and were therefore designated as the Nestow Soil Series. It is likely that both these occur in the sandy landscapes of the Edmonton East grid cell, with the more acidic Nestow series being relatively more sensitive than the Primula Soil Series.

With regard to differences in the H-F map and the ARC model results for the Luvisolic soils of the Islet Upland, the buffering capacity of the litter layer in these soils appears to be a major factor affecting sensitivity. In applying the ARC model, the effect of acid input on the litter was first calculated. The water percolating through the litter was thus reduced in acidity, and the impact on the mineral surface horizon(s) was diminished. The critical loads were based on the chemical effect on the mineral part of the soil (0-25 cm) and not on the litter. Consequently, the Luvisols under native forest in the Edmonton east grid cell were determined to have Low sensitivity to acidification.

6.2 SENSITIVITY MAP

From the results in Table 18, critical loads mapping was based on assignment of the most acid sensitive soil series to categories as follows:

Nestow (Primula)	Moderate to Sensitive
Helliwell	Moderate to Low
Mundare	Moderate to Low

A map depicting the Land Systems, land cover and soil sensitivity to acid inputs in the Edmonton East map sheet was developed based on soils and land cover information as described in Sections 3.1 and 3.3. The proportions of land cover in each Land System under the categories of Cultivated, Trees, Shrubs, Grasslands, Wetlands and Other Lands are provided in Appendix C.

The soil rating for sensitivity to acid inputs, as determined in the previous section, was superimposed on the land cover information. Proportions of soil series within Land Systems were estimated from information provided in Table 2, and from this, the proportions of soils in Moderate to Sensitive (Nestow and Primula), Moderate to Low (Helliwell and Mundare), and Low (all other soils) acidification sensitivity categories were derived. The sensitivity ratings were applied only to soils under grassland, tree, shrub and wetland land cover types. Cultivated soils were not rated, nor were lands categorized as 'Other Lands'.

The sensitivity category proportions in the various Land Systems are given in Table 19. There were five Land Systems that had a component of the Sensitive or Moderate categories indicated above. The derivation of percentages of different sensitivity classes in these five Land Systems is provided in Appendix D.

Land System Symbol	Land System Name	Major Soils	Minor Soils	Acidification Sensitivity
05.00.05	North Saskatchewan River Valley	Eroded with Black Chernozems	Penhold (Orthic Black Chernozem) Gleysols/ Water	Low - 57% Cultivated - 16% Wetland, Water - 27%
05.3d.05:	Redwater Plain	Mundare (Orthic Black Chernozem)	Primula (Eluviated Eutric Brunisol) Peace Hills (Orthic Black Chernozem)	Low - 19% Moderate - 28% Sensitive - 7% Cultivated - 46%
05.3d.09	Partridge Plain	Angus Ridge (Eluviated Black Chernozem)	Rolly View (Orthic Dark Gray Chernozem) Gleysols/ Water	Low - 20% Cultivated - 80%
05.3d.27	Ferlow Plain	Angus Ridge (Eluviated Black Chernozem)	Rolly View (Orthic Dark Gray Chernozem) Angus Ridge (Eluviated Black Chernozem)	Low - 51% Cultivated - 47% Wetland, Water - 2%
05.4a.02	Lonestar Plain	Killam (Black Solodized Solonetz)	Heisler (Solonetzic Black Chernozem) Daysland (Black Solod)	Low - 58% Cultivated - 42%
05.4a.03	Irys Plain	Killam (Black Solodized Solonetz) Shonts (Black Solodized Solonetz)	Gleysols/ Water Daysland	Low - 54% Cultivated - 46%
05.4a.04	Bruce Plain	Killam (Black Solodized Solonetz)	Daysland (Black Solod) Gleysols/ Water	Low - 66% Cultivated - 34%
05.4a.05	Daysland Plain	Heisler (Solonetzic Black Chernozem)	Elnora (Orthic Black Chernozem) Foreman (Solonetzic Humic Gleysol)	Low - 37% Cultivated - 63%
05.4a.06	Bawlf Plain	Killam (Black Solodized Solonetz)	Heisler (Solonetzic Black Chernozem) Gleysols/ Water	Low - 37% Cultivated - 63%
05.4a.14	Little Beaver Plain	Norma (Solonetzic Black Chernozem)	Camrose (Black Solodized Solonetz) Gleysols/ Water	Low - 53% Cultivated - 46% Wetland, Water - 1%

Table 19. Acidification Sensitivity of Land Systems in the Edmonton East Grid Cell

Land System Symbol	Land System Name	Major Soils	Minor Soils	Acidification Sensitivity
05.4a.15	Ryley Plain	Camrose (Black Solodized Solonetz)	Norma (Solonetzic Black Chernozem) Gleysols/ Water	Low - 51% Cultivated - 49%
05.4a.16	Beaverhill Lake	Water	Gleysols/ Water	Low - 37% Cultivated - 5% Wetland, Water - 58%
05.4a.17	Chipman Plain	Camrose (Black Solodized Solonetz) Angus Ridge (Eluviated Black Chernozem)	Gleysols/ Water Kavanagh (Black Solodized Solonetz)	Low - 47% Cultivated - 53%
05.4a.19	Katchemut Upland	Angus Ridge (Eluviated Black Chernozem)	Norma (Solonetzic Black Chernozem) Gleysols/ Water	Low - 26% Cultivated - 74%
05.5b.04	Royal Park Plain	Angus Ridge (Eluviated Black Chernozem) Hobbema (Eluviated Black Chernozem)	Gleysols/ Water	Low - 21% Cultivated - 79%
05.5b.05	Inland Plain	Norma (Solonetzic Black Chernozem) Angus Ridge (Eluviated Black Chernozem) Camrose (Black Solodized Solonetz)	Gleysols/ Water	Low - 31% Cultivated - 69%
05.5b.06	Vegreville Plain	Camrose (Black Solodized Solonetz) Angus Ridge (Eluviated Black Chernozem)	Gleysols/ Water ZSZzbl (Misc. Solonetzic and Black Chernozems)	Low - 50% Cultivated - 50%
05.5b.08:	Whitford Plain	Angus Ridge (Eluviated Black Chernozem)	Ukalta (Orthic Black Chernozem) Navarre (Gleyed Black Chernozem)	Low - 21% Cultivated - 79%
05.5b.09	Norma Plain	Angus Ridge (Eluviated Black Chernozem) Camrose (Black Solodized Solonetz)	Mundare (Orthic Black Chernozem) Peace Hills (Eluviated Black Chernozem)	Low - 17% Moderate - 8% Cultivated - 75%
05.5b.12	Watt Lake Plain	Angus Ridge (Eluviated Black Chernozem) Camrose (Black Solodized Solonetz)	Hairy Hill (Rego Humic Gleysol) Norma (Solonetzic Black Chernozem)	Low - 30% Cultivated - 67% Wetland, Water - 3%
05.5b.13	Hairy Hill Plain	Rolly View (Orthic Dark Gray Chernozem)	Redwater (Orthic Dark Gray Chernozem) Ponoka (Eluviated Black Chernozem)	Cultivated - 81% Other Land - 19%
05.5b.14:	Hilliard Plain	Angus Ridge (Eluviated Black Chernozem)	Camrose (Black Solodized Solonetz) Gleysols/ Water	Low - 25% Cultivated - 75%
05.5b.15:	Kahwin Plain	Angus Ridge (Eluviated Black Chernozem)	Ponoka (Eluviated Black Chernozem) Gleysols/ Water	Low - 15% Cultivated - 85%
05.6.01	Islet Upland	Cooking Lake (Orthic Gray Luvisol) Uncas (Dark Gray Luvisol)	Gleysols/ Water Mico (Orthic Dark Gray Chernozem)	Low - 81% Cultivated - 12% Wetland, Water - 7%

Table 19. Acidification Sensitivity of Land Systems in the Edmonton East Grid Cell

Land System Symbol	Land System Name	Major Soils	Minor Soils	Acidification Sensitivity
06.2a.01	Musidora Upland	Helliwell (Orthic Dark Gray Chernozem) Peace Hills (Orthic Black Chernozem)	Primula (Eluviated Eutric Brunisol) Misc. Organics	Low - 43% Moderate - 25% Sensitive - 8% Cultivated - 24%
06.2a.11	Eldorena Plain	Peace Hills (Orthic Black Chernozem) Primula (Eluviated Eutric Brunisol)	Manatokan (Terric Mesisol) Mundare (Eluviated Black Chernozem)	Low - 34% Moderate - 24% Sensitive - 17% Cultivated - 25%
06.2b.05	Edward Upland	Primula (Eluviated Eutric Brunisol) Redwater (Orthic Dark Gray Chernozem)	Uncas (Dark Gray Luvisol) Misc. Organics	Low - 64% Moderate - 17% Sensitive - 17% Cultivated - 2%
06.2b.08	Redclay Plain	Angus Ridge (Eluviated Black Chernozem) Ponoka (Eluviated Black Chernozem)	Rimbey (Orthic Dark Gray Chernozem) Rolly View (Orthic Dark Gray Chernozem)	Low - 29% Cultivated - 70% Wetland, Water - 1%
06.2b.12	Delph Upland	Angus Ridge (Eluviated Black Chernozem) Uncas (Dark Gray Luvisol)	Rolly View (Orthic Dark Gray Chernozem) Gleysols/ Water	Low - 21% Cultivated - 79%
06.2b.17	Pakan Plain	Ponoka (Eluviated Black Chernozem)	Hobbema (Eluviated Black Chernozem) Kavanagh (Black Solodized Solonetz)	Low - 28% Cultivated - 72%
06.2c.25	Thorhild Plain	Spedden (Dark Gray Luvisol) Kehiwin (Dark Gray Chernozem)	LaCorey (Orthic Gray Luvisol) Gleysols/ Water	Low - 39% Cultivated - 61%

Table 19.	Acidification Sensitivit	v of Land Svs	tems in the Edmonto	n East Grid Cell
		, eaa e,e		

For purposes of developing a soil sensitivity map, the Land Systems with Sensitive and/or Moderate inclusions were grouped together such that there were three Sensitivity categories on the map. These generalized categories, and the Land Systems in the categories, are indicated in Table 20.

Acidification Sensitivity Category	Land System
	Musidora Upland,
Low - Moderate - Sensitive Mix	Redwater Plain,
	Eldorena Plain,
	Edward Upland
Low - Moderate Mix	Norma Plain
Low	All other Land Systems

The above sensitivity categories were identified by colour coding on the critical loads map. The information from Table 19 is presented as a legend on the map. The map is provided on a CD as well as in hard copy form in the back pocket of this report.

7.0 CONCLUSIONS

The objective of this study was to derive a critical load for the Edmonton East map sheet, an area identified as possibly having significant areas of sensitive and moderate sensitivity soils. Critical loads as low as 0.15 kmol H⁺ ha⁻¹ yr⁻¹ over a 50 year assessment period, and as low as 0.1 kmol H⁺ ha⁻¹ yr⁻¹ over a 100 year assessment period were estimated by application of the ARC soil acidification model. Most of the soils were determined to have critical loads greater than 1.0 kmol H⁺ ha⁻¹ yr⁻¹. Critical loads determined by two other methods, namely the empirical Skokloster method and the Steady State Mass Balance method, both based mainly on weathering estimates, were in general agreement with critical loads based on the ARC model.

Portions of five land systems in the Edmonton 83H East Half grid cell were characterized as having a component of Sensitive and Moderately Sensitive soils. These are the Eldorena Plain, Redwater Plain, Musidora Upland, Edward Upland and Norma Plain. Other Land Systems likely have small components of Sensitive and Moderately Sensitive soils, but of too low extent to enable mapping at the scale applied in this assessment. The assignment of Sensitive, Moderate and Low Sensitivity categories was applied only to lands classified as having grassland, tree or shrub cover, on the basis of land use mapping by the Prairie Farm Rehabilitation Administration in 1993-1995. Sensitive soils account for 0.5% and Moderately Sensitive soils account for 1.3% of the entire grid cell area. As defined in the Acid Deposition Management Framework (Clean Air Strategic Alliance and Alberta Environment 1999), this finding does not support the assignment of this grid cell to a Sensitive or Moderate Sensitivity rating.

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APPENDICES

APPENDIX A: SOIL PROFILE DESCRIPTIONS

Site 1

Location: Classification:	LSD 16 - Section 7 - Township 55 - Range 14 - West 4 Meridian
Subgroup:	Eluviated Black Chernozem
Series:	Mundare – Eluviated Variant
Landform:	
Genetic Material:	Glaciofluvial
Surface Expression:	Undulating; 6-9% slopes
Drainage/ Perviousness:	Rapidly drained; high perviousness
Site Features:	South-west slope; mid slope position; non-stony
Vegetation:	Native and introduced grasses, shrubs, aspen in hollows

Profile Description:

Ah	0 to 28 cm	Black (10YR 3/1 dry); loamy sand; weak granular; no coarse fragments.
Ae	28 to 35 cm	Gray; loamy sand; weak, medium platy; no coarse fragments.
Bm	35 to 1 m	Brown to dark brown; loamy sand; single grain; loose; 2% coarse fragments.

Site 2

Location: Classification:	LSD 12 - Section 14 - Township 56 - Range 16 - West 4 Meridian
Subgroup: Series:	Orthic Black Chernozem; possibly Solonetzic Black Chernozem Ukalta
Landform:	
Genetic Material:	Glaciofluvial over till
Surface Expression:	Undulating to hummocky; 6-9% slopes
Drainage/ Perviousness:	Rapidly drained; high perviousness
Site Features:	Located in shallow draw down slope from Site 1; toe slope position; 5- West aspect; mid slope position; non-stony
Vegetation:	Grasses, wild rose, buckbrush,

Profile Description:

Ah	0 to 25 cm	Black (10YR 2/1); loamy sand - sandy loam; weak granular
Ae	25 to 31 cm	Gray; loamy sand - sandy loam; weak platy; no coarse fragments; gradual,
Btnj	31 to 50 cm	smooth boundary. Dark gray; clay loam; weak, medium columnar breaking to strong, medium, subangular blocky; hard; few coarse fragments.

Site 3

Location: Classification:	LSD 16 - Section 5 - Township 58 - Range 14 - West 4 Meridian
Subgroup:	Eluviated Dystric Brunisol
Series:	Nestow; variant with relatively thick Ahe
Landform:	Nestow, variant with relatively thick Ane
Genetic Material:	Glaciofluvial
Surface Expression:	Undulating; 5% slopes
Drainage/ Perviousness:	Rapidly drained; high perviousness
Site Features:	Southeast aspect; mid position; non-stony
Vegetation:	Jackpine, bearberry, grasses

Profile Description:

LFH	5 to 0 cm	Dark brown to black; needle-grass litter
Ahe1	0 to 7 cm	Dark grayish brown (10YR 4/1 dry); sand; single grain; loose; abundant, fine to coarse roots; no coarse fragments.
Ahe2	7 to 14 cm	Dark grayish brown (10YR 4/3 dry); sand; single grain; loose; abundant, fine to medium roots; no coarse fragments.
Bm	14 to 26	Brown (7.5YR 5/4 dry); sand; single grain; loose; few, fine roots; no coarse fragments.
	cm	
Btgj	26 to 34 cm	Brown (10YR 5/3 dry); sandy loam (gravelly) ; very weak subangular blocky; very friable; very few, fine roots; 20% gravelly coarse fragments.
Bm	34 to 100 cm	Yellowish brown (10YR 5/4 dry); sand; single grain; loose; no roots; no coarse fragments.

Site 4

Location: Classification:	LSD 13 - Section 3 - Township 58 - Range 14 - West 4 Meridian
Subgroup: Series:	Eluviated Dystric Brunisol Nestow
Landform:	
Genetic Material:	Glaciofluvial
Surface Expression:	Undulating; 2-5% slopes
Drainage/ Perviousness:	Rapidly drained; high perviousness
Site Features:	Almost level sample site; non-stony
Vegetation:	Jackpine, bearberry, blueberry, grasses

Profile Description:

LFH	7 to 0 cm	Dark brown to black; needle-grass litter
Ahe	0 to 2 cm	Dark gray (10YR 4.5/1 dry); sand; single grain; loose; abundant, fine to coarse
		roots; no coarse fragments.
Ae	2 to 12 cm	Gray (10YR 5.5/3 dry); sand; single grain; loose; abundant, fine to medium
		roots; no coarse fragments.
Bm	12 to 25+ cm	Pale brown (10YR 6/3 dry); sand; single grain; loose; few, fine roots; no coarse
		fragments.

Profile Description:

LFH	1 to 0 cm	Dark brown to black; needle-lichen litter
Ae	0 to 6 cm	Light brownish gray (10YR 6/2 dry); medium sand; single grain; loose; plentiful,
		fine to coarse roots; no coarse fragments.
Bm1	6 to 28 cm	Light yellowish brown (10YR 6/4 dry); medium sand; single grain; loose; few,
		fine roots; no coarse fragments.
Bm2	28 to 80 cm	Pale brown (10YR 6/3 dry); medium sand; single grain; loose; very few roots; no
		coarse fragments.
BC	80 to 100+ cm	Light yellowish brown (10YR 6/4 dry); medium sand; single grain; loose; no
		roots; no coarse fragments.

Site 6

Location: Classification:	LSD 16 - Section 20 - Township 57 - Range 20 - West 4 Meridian
Subgroup:	Eluviated Dystric Brunisol
Series: Landform:	Nestow
Genetic Material:	Eolian
Surface Expression:	Undulating, ridged and hummocky (duned); 6-9% slopes, some 10-15%
Drainage/ Perviousness:	Rapidly drained; high perviousness
Site Features: Vegetation:	West aspect; 9% slope; upper slope position; non-stony Jackpine, bearberry, grasses

LFH	3 to 0 cm	Dark brown to black; needle litter
Ae	0 to 10 cm	Light gray (10YR 7/1 dry); medium sand; single grain; loose; plentiful, fine to
		coarse roots; no coarse fragments.
Bm1	10 to 30 cm	Yellowish brown (10YR 5.5/4 dry); medium sand; single grain; loose; few, fine
		roots; no coarse fragments.
Bm2	30 to 80 cm	Pale brown (10YR 6/3 dry); medium sand; single grain; loose; very few roots; no
		coarse fragments.
BC	80 to 100+ cm	Light yellowish brown (10YR 6/4 dry); medium sand; single grain; loose; no roots; no coarse fragments.

Location: Classification:	LSD 1 - Section 30 - Township 56 - Range 20 - West 4 Meridian
Subgroup: Series:	Eluviated Dystric Brunisol Nestow
Landform:	Nestow
Genetic Material:	Eolian
Surface Expression:	Undulating; 2-5% slopes
Drainage/ Perviousness:	Rapidly drained; high perviousness
Site Features: Vegetation:	West aspect; 3% slope; mid slope position; non-stony Jackpine, shrubs, grasses; some aspen nearby

Profile Description:

2 to 0 cm	Dark brown to black; leaf and needle litter
0 to 11 cm	Light gray (10YR 7/1 dry); medium sand; single grain; loose; abundant, fine to coarse roots; no coarse fragments.
11 to 30 cm	Yellowish brown (10YR 5.5/4 dry); medium sand; single grain; loose; few, fine
	roots; no coarse fragments.
30 to 80 cm	Pale brown (10YR 6/3 dry); medium sand; single grain; loose; very few roots; no
	coarse fragments.
80 to 100+ cm	Brown (10YR 5/3 dry); medium sand; single grain; loose; no roots; no coarse fragments.
	0 to 11 cm 11 to 30 cm 30 to 80 cm

Site 8

Location: Classification:	LSD 4 - Section 28 - Township 56 - Range 20 - West 4 Meridian
Subgroup:	Eluviated Dystric Brunisol
Series:	Nestow
Landform:	
Genetic Material:	Eolian
Surface Expression:	Undulating and hummocky (duned); 6-9% slopes, some 10-15%
Drainage/ Perviousness:	Rapidly drained; high perviousness
Site Features:	West aspect; 10% slope; upper slope position; non-stony
Vegetation:	Jackpine, bearberry; aspen down slope

LFH	2 to 0 cm	Dark brown to black; needle and leaf litter
Ahe	0 to 4 cm	Grayish brown (10YR 5/2 dry); medium sand; single grain; loose; abundant, fine
		to coarse roots; no coarse fragments.
Ae	4 to 12 cm	Gray (10YR 6/1 dry); medium sand; single grain; loose; plentiful, fine roots; no coarse fragments.
Bm1	12 to 65 cm	Yellowish brown (10YR 5/4 dry); medium sand; single grain; loose; very few
		roots; no coarse fragments.
Bm2	65 to 100+ cm	Brown (10YR 5/3 dry); medium sand; single grain; loose; no roots; no coarse fragments.

Location: Classification:	LSD 8 - Section 34 - Township 57 - Range 17 - West 4 Meridian
Subgroup: Series:	Eluviated Dystric Brunisol Nestow
Landform:	Nestow
Genetic Material:	Eolian
Surface Expression:	Undulating; 6-9% slopes
Drainage/ Perviousness:	Rapidly drained; high perviousness
Site Features: Vegetation:	Northwest aspect; 9% slope; mid slope position; non-stony Jackpine, bearberry, lichen

Profile Description:

LFH	1 to 0 cm	Dark brown; needle-lichen litter
Ae	0 to 6 cm	Light brownish gray (10YR 5.5/2 dry); medium sand; single grain; loose; abundant, fine to coarse roots; no coarse fragments.
Bm1	6 to 35 cm	Pale brown (10YR 6/3 dry); medium sand; single grain; loose; few, fine roots; no coarse fragments.
Bm2	35 to 90 cm	Light yellowish brown (10YR 6/4 dry); medium sand; single grain; loose; very few roots; no coarse fragments.
BC	90+ cm	Grayish brown (2.5Y 5/2 dry); medium sand; single grain; loose; no roots; no coarse fragments.

Site 10

Location: Classification:	LSD 1 - Section 25 - Township 57- Range 18 - West 4 Meridian
Subgroup: Series:	Eluviated Dystric Brunisol Nestow
Landform:	
Genetic Material:	Eolian
Surface Expression:	Undulating; 2-5% slopes
Drainage/ Perviousness:	Rapidly drained; high perviousness
Site Features: Vegetation:	West aspect of 5% slope; mid slope position; non stony Jackpine, bearberry, lichen, minor grasses

LFH	(1-2) to 0 cm	Dark brown to black; needle-lichen litter
Ae	0 to 7 cm	Light gray (10YR 6.5/1 dry); medium sand; single grain; loose; abundant, fine to
		coarse roots; no coarse fragments.
Bm1	7 to 35 cm	Yellowish brown (10YR 5/4 dry); medium sand; single grain; loose; few, fine roots; no coarse fragments.
Bm2	35 to 80 cm	Pale brown (10YR 6/3 dry); medium sand; single grain; loose; very few roots; no
		coarse fragments.
BC	80 to 100+ cm	Brown (10YR 5/3 dry); medium sand; single grain; loose; no roots; no coarse fragments.

Location: Classification:	LSD 5 - Section 27 - Township 49 - Range 20 - West 4 Meridian
Subgroup:	Dark Gray Luvisol
Series:	Uncas
Landform	
Genetic Material:	Glacial till
Surface Expression:	Hummocky; 6-9% slopes; some 10-15%
Drainage/ Perviousness:	Well drained; medium perviousness
Site Features:	West aspect of 7-9% slope; mid slope position; slightly stony
Vegetation:	Brome grass

Profile Description:

Ар	0 to 18 cm	Dark gray (10YR 4/1 dry); sandy loam; mix of Ae and Ah/Ahe material; few coarse fragments.
Bt1	18 to 42 cm	Dark yellowish brown (10YR 4/4 dry); sandy clay loam; moderate, medium subangular blocky; friable to firm; few coarse fragments.
Bt2	42 to 70 cm	Yellowish brown (10YR 4-5/4 dry); yellower than Bt1; sandy clay loam; weak, medium subangular blocky; friable; few coarse fragments.
BC	70 to 100+ cm	Gray (2.5Y 5/1 dry); sandy clay loam - clay loam; massive; friable to firm; few coarse fragments.

Site 12

Location: Classification:	LSD 12 - Section 11 - Township 51 - Range 20 - West 4 Meridian
Subgroup:	Dark Gray Luvisol
Series:	Uncas
Landform:	
Genetic Material:	Glacial till
Surface Expression:	Undulating; 6-9% slopes
Drainage/ Perviousness:	Well drained; Medium perviousness
Site Features:	Northeast aspect of 6% slope; mid slope position; slightly stony
Vegetation:	Aspen, dogwood, wild rose, alder, hazelnut

LFH Ahe	13 to 0 cm 0 to 5 cm	Dark brown leaf litter; matted with abundant fine roots near base Dark gray (10YR 4.5-5/1 dry); sandy loam; weak platy; abundant fine to coarse roots; few coarse fragments.
Ae	5 to 16 cm	Gray (10YR 5.5/1 dry); sandy loam; moderate, medium platy; plentiful fine to coarse roots; few coarse fragments.
AB	16 to 28 cm	Grayish brown (10YR 5/2 dry); sandy clay loam; weak, medium subangular blocky; friable to firm; few coarse fragments.
Bt	28 to 65 cm	Dark yellowish brown (10YR 4/4 dry); clay loam; moderate, medium subangular blocky; firm; few coarse fragments.
BC	65 to 100+ cm	Brown (10YR 4/3 dry); clay loam; massive; firm; few coarse fragments.

Location: Classification:	LSD 4 - Section 6 - Township 52 - Range 19 - West 4 Meridian
Subgroup:	Orthic Gray Luvisol
Series:	Cooking Lake
Landform:	U U U U U U U U U U U U U U U U U U U
Genetic Material:	Glacial till
Surface Expression:	Hummocky; 10-15% slopes
Drainage/ Perviousness:	Well drained; medium perviousness
Site Features:	East aspect of 15% slope; mid slope position; slightly stony;
Vegetation:	Aspen, dogwood, wild rose, alder, hazelnut

Profile Description:

LFH Ae	12 to 0 cm 0 to 13 cm	Dark brown leaf litter; matted with abundant fine roots near base Light gray (10YR 7/1 dry); sandy loam; moderate, medium platy; abundant fine to coarse roots; few coarse fragments.
Bt1	13 to 40 cm	Brown (10YR 6/2 & 5/3 dry); clay loam; moderate, medium subangular blocky; friable to firm; few roots; few coarse fragments.
Bt2	40 to 65 cm	Yellowish brown (10YR 5/4 dry); clay loam; moderate, medium subangular blocky; firm; very few roots; few coarse fragments.
BC	65 to 100+ cm	Brown (10YR 5/3 dry); clay loam; massive; firm; few coarse fragments.

Site 15*

Location: Classification:	LSD 11 - Section 11 - Township 57 - Range 17 - West 4 Meridian
Subgroup:	Orthic Black Chernozem
Series: Landform:	Peace Hills
Genetic Material:	Thin eolian over glaciofluvial
Surface Expression:	Undulating; 2-5% slopes
Drainage/ Perviousness:	Rapidly drained; high perviousness
Site Features:	West aspect; upper position 4% slope; non-stony
Vegetation:	Cultivated; grasses

Profile Description:

Ap	0 to 15 cm	Very dark gray (10YR 3.5/1); loamy sand to sandy loam; weak granular; plentiful
		fine roots
AB	15 to 30 cm	Brownish gray; loamy sand to sandy loam
Bm	30-80	Reddish brown; loamy sand to sandy loam; single grain; very few roots
BC	80 to 100 cm	Grayish brown; loamy sand to sandy loam; single grain; loose.

(* Note: There is no Site 14.)

Location: Classification:	LSD 11 - Section 33 - Township 52 - Range 19 - West 4 Meridian
Subgroup:	Orthic Gray Luvisol
Series: Landform:	Cooking Lake
Genetic Material: Surface Expression:	Glacial till Hummocky; 10-15% slopes
Drainage/ Perviousness: Site Features: Vegetation:	Well drained; medium perviousness East aspect of 13-15% slope; upper slope position; slightly stony Aspen, wild rose, alder, hazelnut, grasses

Profile Description:

LFH Ahe	9 to 0 cm 0 to 4 cm	Dark brown leaf litter; matted with abundant fine roots near base Gray (10YR 5/1 dry); sandy loam; weak, medium platy to granular; abundant fine
And	0 10 4 011	to coarse roots; few coarse fragments.
Ae1	4 to 18 cm	Light brownish gray (10YR 6/2 dry); sandy loam; moderate, medium platy; plentiful fine to coarse roots; few coarse fragments.
Ae2	18 to 35 cm	Light gray (10YR 6.5/2 dry); sandy loam; moderate, medium platy; few fine to coarse roots; few coarse fragments.
Bt	35 to 60+ cm	Brown (10YR 4/3 dry); clay loam; moderate, medium subangular blocky; friable to firm; few roots; few coarse fragments.

SITE 17

Location:	LSD 16 - Section 28 - Township 53 - Range 20 - West 4 Meridian; near Tawayik Lake, Elk Island National Park
Classification:	
Subgroup:	Orthic Gray Luvisol
Series:	Cooking Lake
Landform:	ů –
Genetic Material:	Glacial till; thin overlay of non-stony material
Surface Expression:	Undulating; 6-9% slopes; about 100 m from slope down to Tawayik Lake
Drainage/ Perviousness:	Moderately well drained; medium to low perviousness
Site Features:	East aspect; 3% slope; upper slope position; non stony
Vegetation:	Aspen, wild rose, hazelnut, grasses

LFH	7 to 0 cm	Dark brown leaf litter
Ahe	0 to 3 cm	Dark gray (10YR 4.5/1 dry); sandy loam; weak, medium platy to granular; abundant fine to coarse roots; no coarse fragments.
Ae1	3 to 12 cm	Gray (10YR 6/1 dry); sandy loam; moderate, medium platy; plentiful fine to coarse roots; no coarse fragments.
Ae2	12 to 26 cm	Light gray (10YR 7/1 dry); sandy loam; moderate, medium platy; few fine to coarse roots; few coarse fragments.
Bt1	26 to 40 cm	Dark yellowish brown (10YR 4/4 dry); clay loam; strong, medium subangular blocky; friable to firm; few roots; few coarse fragments.
Bt2	40 to 50 cm	Brown (10YR 5/3 dry); clay loam; moderate, medium subangular blocky; friable to firm; few roots; few coarse fragments.
Bt3	50 to 65+ cm	Brown (10YR 5/3 dry); clay loam; weak, medium subangular blocky; friable to firm; few roots; few coarse fragments.

Location: Classification:	LSD 12 - Section 31 - Township 54 - Range 19 - West 4 Meridian
Subgroup: Series:	Orthic Gray Luvisol Cooking Lake
Landform:	U U U U U U U U U U U U U U U U U U U
Genetic Material: Surface Expression:	Glacial till; Hummocky; 10-15% slopes; near Beaver Pond Parking Area, Elk Island
Drainage/ Perviousness: Site Features: Vegetation:	National Park Moderately well drained; medium perviousness Northeast aspect; 12-15% slope; mid slope position; slightly stony Aspen, wild rose, hazelnut, wintergreen

Profile Description:

LFH	7 to 0 cm	Dark brown leaf litter
Ahe	0 to 5 cm	Dark gray (10YR 4/1 with 6/1 dry); sandy loam; weak, medium platy to granular; abundant fine to coarse roots; few coarse fragments.
Ae	5 to 21 cm	Gray (10YR 6/2 dry); sandy loam; moderate, medium platy; plentiful fine to coarse roots; few coarse fragments.
Bt1	21 to 55 cm	Dark yellowish brown (10YR 4/4 dry); clay loam; strong, medium subangular blocky; friable to firm; few roots; few coarse fragments.
BC	55 to 80+ cm	Brown (10YR 4-5/3 dry); clay loam; massive; firm; few roots; few coarse fragments.

SITE 19

Location:	LSD 8 - Section 31 - Township 54 - Range 20 - West 4 Meridian
Classification: Subgroup:	Orthic Dark Gray Chernozem
Series:	Helliwell – gravelly variant
Landform:	
Genetic Material:	Glaciofluvial
Surface Expression:	Undulating; 2-5% slopes; some 6-9%;
Drainage/ Perviousness:	Rapidly drained; high perviousness
Site Features:	East aspect; 4-5% slope; mid slope position; slightly stony
Vegetation:	Jackpine, aspen, grasses, some moss, wild rose

LFH	6 to 0 cm	Dark brown needle and leaf litter
Ahe1	0 to 6 cm	Dark gray (10YR 4/1); loamy sand; single grain; loose; abundant fine to coarse roots; few, gravelly coarse fragments.
Ahe2	6 to 20 cm	Brown (10YR 4-5/3 dry); sand; single grain; loose; plentiful fine to coarse roots; few coarse fragments.
Bm1	20 to 35 cm	Dark yellowish brown (10YR 4/4 dry); sand; single grain; loose; few roots; few coarse fragments.
Bm2	35 to 60 cm	Brown (10YR 5/3 dry); sand (gravelly); single grain; loose; very few roots; ~20% gravelly coarse fragments.
Bm3	60 to 100 cm	Brown (10YR 4.5/3 dry); sand (gravelly); single grain; loose; very few roots; ~20% gravelly coarse fragments.

Location: Classification:	LSD 2 - Section 29 - Township 54 - Range 14 - West 4 Meridian
Subgroup:	Orthic Black Chernozem
Series:	Ukalta
Landform:	
Genetic Material:	Glaciofluvial over till
Surface Expression:	Undulating; 6-9% slopes
Drainage/ Perviousness:	Rapidly drained; high perviousness
Site Features:	Northeast aspect; 6-7% slope; mid slope position; non stony
Vegetation:	Native pasture; aspen, grasses, wild rose, other shrubs

Profile Description:

LFH Ah	3 to 0 cm 0 to 23 cm	Dark brown leaf litter Black (10YR 2.5/1); sandy loam; weak granular; very friable; abundant fine to coarse roots; no coarse fragments.
		0
Bm1	23 to 40 cm	Yellowish brown (10YR 5/4 dry); sandy loam; weak prismatic; very friable; few
		roots; no coarse fragments.
Bm2	40 to 65 cm	Brown (10YR 5/3 dry); sandy loam; weak prismatic; very friable; very few roots;
		no coarse fragments.
IIBC	65 to 100 cm	Brown (10YR 4/3 dry); clay loam; massive; friable; no roots; few coarse
		fragments.

SITE 21

Location: Classification:	LSD 8 - Section 5 - Township 57 - Range 14 - West 4 Meridian
Subgroup: Series:	Orthic Dark Gray Chernozem Redwater
Landform:	
Genetic Material:	Glaciofluvial
Surface Expression:	Undulating; 2-5% slopes
Drainage/ Perviousness:	Rapidly drained; high perviousness
Site Features: Vegetation:	South aspect; 4% slope; mid slope position; non stony Native pasture; jackpine, shrubs, aspen, grasses, mosses

LFH	5 to 0 cm	Dark brown needle and leaf litter
Ahe	0 to 26 cm	Dark gray (10YR 4/1); sandy loam; weak granular; very friable; abundant fine to coarse roots; no coarse fragments.
Bm1	26 to 40 cm	Yellowish brown (10YR 5/4 dry); sandy loam; weak prismatic; very friable; few roots; no coarse fragments.
Bm2	40 to 65 cm	Brown (10YR 5/3 dry); sandy loam; weak prismatic; very friable; very few roots;
50		no coarse fragments.
BC	65 to 100 cm	Brown (10YR 4/3 dry); sandy loam; massive; friable; no roots; few coarse fragments.

Location: Classification:	LSD 12 - Section 28 - Township 57 - Range 18 - West 4 Meridian
Subgroup:	Orthic Black Chernozem
Series:	Helliwell – gravelly variant
Landform:	
Genetic Material:	Glaciofluvial
Surface Expression:	Undulating; 2-5% slopes
Drainage/ Perviousness:	Rapidly drained; high perviousness
Site Features:	East aspect; 2% slope; mid slope position; slightly stony; near gravel pit.
Vegetation:	Forage; grasses; shrubs returning in spots

Profile Description:

Ap Btj	0 to 9 cm 9 to 18 cm	Black (10YR 3.5/2 dry); loamy sand; single grain; loose; few coarse fragments. Dark yellowish brown (10YR 4/4); loamy sand; weak, medium subangular blocky; very friable; few fine to coarse roots; 5% gravelly coarse fragments.
Bm1	18 to 35 cm	Brown (10YR 4/3 dry); loamy sand; single grain; loose; very few roots; 5-10% gravelly coarse fragments.
Bm2	35 to 60 cm	Brown (10YR 4-5/3 dry); loamy sand; single grain; loose; very few roots; 5-10% gravelly coarse fragments.
Bm3	60 to 80 cm	Brown (10YR 5/3 dry); loamy sand; single grain; loose; very few roots; 20-30% gravelly coarse fragments.
BC	80 to 100 cm	Grayish brown (10YR 5/2 dry); sandy loam; massive; friable; no roots; 20-30% gravelly coarse fragments.

SITE 23

Location: Classification:	LSD 2 - Section 29 - Township 57 - Range 18 - West 4 Meridian
Subgroup:	Eluviated Dystric Brunisol
Series: Landform:	Nestow
Genetic Material: Surface Expression:	Eolian Hummocky; 10-15% slopes; some 6-9%
Drainage/ Perviousness:	Rapidly drained; high perviousness
Site Features: Vegetation:	Northeast aspect; 10-12% slope; mid slope position; non stony; Jackpine, lichen, feathermoss

LF	4 to 0 cm	Dark brown; needle, moss, lichen litter
Ae	0 to 8 cm	Gray (10YR 5/2 dry); sand; single grain; loose; plentiful, fine to coarse roots; no
		coarse fragments.
Bm1	8 to 40 cm	Pale brown (10YR 6/3 dry); sand; single grain; loose; few roots; no coarse
		fragments.
Bm2	40 to 70 cm	Brown (10YR 5/3 dry); sand; single grain; loose; no roots; no coarse fragments.
Bm3	70 to 90 cm	Brown (10YR 4-5/3 dry); sand; single grain; loose; no roots; no coarse fragments.
BC	90 to 100+ cm	Brown (10YR 5/3 dry); sand; single grain; loose; no coarse fragments.

Location: Classification:	LSD 3 - Section 22 - Township 57 - Range 19 - West 4 Meridian
Subgroup: Series:	Orthic Dark Gray Chernozem Helliwell
Landform:	
Genetic Material: Surface Expression:	Glaciofluvial
Drainage/ Perviousness:	Undulating; 2-5% slopes Rapidly drained; high perviousness; may be imperfectly to poorly drained below 1 m
Site Features: Vegetation: Notes:	South aspect; 3-4% slope; lower slope position; non stony; Native pasture; aspen, shrubs, moss, grass Very thick Ae or AB horizon

Profile Description:

LF	4 to 9 cm	Dark brown to black; leaf-moss litter
Ahe	0 to 6 cm	Dark gray (10YR 4.5/1 dry); loamy sand; single grain to weak granular; loose; plentiful, fine to coarse roots; no coarse fragments.
Ae1	6 to 16 cm	Gray (10YR 6/1 dry); loamy sand; single grain; loose; few, fine to coarse roots; no coarse fragments.
Ae2	16 to 80 cm	Light gray (10YR 6-7/1 dry); loamy sand; single grain; loose; very few roots; no coarse fragments.
Bm	80 to 100+ cm	Strong brown (7.5YR 5/6 dry); loamy sand; single grain; loose; no roots; no coarse fragments.

SITE 25

Location: Classification:	LSD 5 - Section 4 - Township 58 - Range 19 - West 4 Meridian
Subgroup:	Orthic Dark Gray Chernozem
Series: Landform:	Helliwell
Genetic Material:	Glaciofluvial; thin eolian overlay
Surface Expression:	Hummocky; 6-9% slopes
Drainage/ Perviousness:	Rapidly drained; high perviousness
Site Features:	Southeast aspect; 9% slope; upper slope position; non stony;
Vegetation:	Forage; grasses

Ар	0 to 15 cm	Dark gray (10YR 4.5/2 dry); loamy sand; single grain; loose; plentiful, fine roots; no coarse fragments.
Ahe	15 to 28 cm	Gray (10YR 5/1 dry); loamy sand; single grain; loose; few, fine roots; no coarse fragments.
Bm1	28 to 55 cm	Brown (10YR 5/3 dry); loamy sand; single grain; loose; no roots; no coarse fragments.
Bm2	55 to 80 cm	Brown (10YR 4-5/3 dry); loamy sand; single grain; loose; no roots; no coarse fragments.
BC	80 to 100+ cm	Brown (10YR 5/3 dry); loamy sand; single grain; loose; no coarse fragments.

Location: Classification:	LSD 13 - Section 15 - Township 57 - Range 17 - West 4 Meridian
Subgroup:	Eluviated Dystric Brunisol
Series:	Nestow
Landform:	
Genetic Material:	Eolian
Surface Expression:	Undulating; 2-5% slopes
Drainage/ Perviousness:	Rapidly drained; high perviousness
Site Features:	East aspect; 4% slope; mid slope position; non stony
Vegetation:	Jackpine, lichen, feather moss

Profile Description:

LF Ae	5 to 0 cm 0 to 4 cm	Dark brown; needle-moss-lichen litter Gray (10YR 5/1 dry); sand; single grain; loose; plentiful, fine to coarse roots; no coarse fragments.
Bm1	4 to 32 cm	Brown (7.5YR 5/4 dry); sand; single grain; loose; few, fine roots; no coarse fragments.
Bm2	32 to 70 cm	Brown (10YR 5/3 dry); sand; single grain; loose; very few roots; no coarse fragments.
Bm3	70 to 100+ cm	Yellowish brown (10YR 5/4 dry); sand; single grain; loose; no roots; no coarse fragments.

Site TH1

TH1 is a soil profile described in AGRASID and in "Soil Survey of Two Hills County N0. 21 Alberta" (Macyk et al. 1985). The profile is referred to as a Peace Hills soil in the Two Hills soil survey, but subsequent adjustment of soil names now places this soil profile in the Mundare soil series.

Location:	SW10 - Township 55 - Range 15 - West 4 Meridian
Classification:	Orthis Block Charnesson
Subgroup: Series:	Orthic Black Chernozem Mundare
Landform:	
Genetic Material:	Glaciofluvial
Surface Expression:	Undulating
Drainage/ Perviousness:	Well drained
Site Features:	N/A
Vegetation	The soil profile is overlain by some drift.

Ah	0 to 25 cm	Very dark gray (10YR 3/1 moist); sand; moderate, medium granular; friable; common roots.
AB1	25 to 35 cm	Dark brown (10YR 3/2 moist); loamy sand; weak, medium granular; friable; few roots.
AB2	35 to 52 cm	Brown to dark brown (7.5YR 4/2 moist); loamy sand to sand; weak, medium granular; friable; few vertical roots.
Bm C	52 to 88 cm 88+ cm	Yellowish brown (10YR 5/4 moist); sand; single grain; loose; few roots. Brown (10YR 5/3 moist); sand; single grain; loose; few roots.

Site BR

BR refers to the soil at the Long-Term Soil Acidification Monitoring site located near Bruderheim, Alberta. The following is the baseline description for this site from Roberts et al. (1989).

Su Se Landf Ge Su Draina	ification: ubgroup: eries: orm: enetic Material: urface Expression age/ Perviousne eatures:						
<u>Profile</u>	e Description:						
LFH	2-0 cm						
Ahe 0-5 cm Very dark grayish brown (10YR 3/2 m); fine fire darkened wood particles; loam; sand to sand; very weak fine granular to single grain; loose; clear wavy boundary; very strongly acid							
Ae (AE	3)5-8 cm	Dark brown to brown (10YR 4/3 m); sand; single grain; loose; clear wavy boundary; very strongly acid					
Bm₁	8-20 cm	Dark yellowish brown (10YR 4/4 m); sand; single grain; loose; clear wavy boundary; very strongly acid					
Bm_2	20-40 cm	Dark yellowish brown to yellowish brown (10YR 4.5/4 m); sand; single grain; loose; gradual wavy boundary; very strongly acid					
Bm₃	40-60 cm Yellowish brown (10YR 5/4 m); sand; single grain; loose; gradual wavy boun strongly acid						
BC	60-80 cm	Yellowish brown (10YR 5/4 to 5.5/4 m); sand; single grain; loose; gradual wavy boundary; strongly acid					
С	80-100+ cm	Light olive yellow (2.5Y 5/4 m); sand; single grain; loose;					

APPENDIX B: SOIL CHEMICAL DATA

Site	Classification	Soil Sorios	Horizon	Depth	рН	Exchangeable Cations and Cation Exchange Capacity (cmol kg ⁻¹)									
Site	Classification	Soli Series	Horizon	(cm)	(CaCl2)						<u> </u>		0-07		Saturation
50.4				0.05		Na	K	Ca	Mg	AI	Fe	Mn	CEC ^z	BC	0.00
ED01	E.BL	Mundare	Ар	0-25	5.9	0.01	0.56	18.35	2.32	< 0.03	< 0.01	0.03	21.8	21.2	0.98
ED02	O.BL	Ukalta	Ah	0-25	5.7	0.10	0.85	28.40	4.82	< 0.03	< 0.01	0.04	39.8	34.2	0.86
5000	0.00		LF	7-0	5.2	< 0.01	1.83	62.96	6.47	0.08	0.06	1.85	88.2	71.3	0.81
ED03	O.DG	Nestow-AA	Ah/Ahe/Ae	0-14	5.3	< 0.01	0.21	15.47	1.10	< 0.03	< 0.01	0.26	21.8	16.8	0.77
			Bm	14-25	4.9	0.01	0.06	2.16	0.30	< 0.03	< 0.01	< 0.02	4.8	2.5	0.53
5004			LF	7-0	4.8	< 0.01	1.59	39.96	5.69	0.08	0.05	1.93	71.4	47.2	0.66
ED04	E.DYB	Nestow-AA	Ah/Ahe	0-12	4.9	0.01	0.15	6.49	0.72	< 0.03	< 0.01	0.20	11.6	7.4	0.64
			Bm	12-25	5.2	<0.01	0.08	1.99	0.25	< 0.03	< 0.01	0.02	3.7	2.3	0.62
	E.DYB	Nestow-AA	LF	1-0	4.8	<0.01	1.62	36.43	4.53	< 0.03	< 0.01	0.84	73.2	42.6	0.58
ED05			Ae	0-6	4.4	<0.01	0.09	1.35	0.25	< 0.03	<0.01	0.06	5.1	1.7	0.33
			Bm	6-25	4.5	<0.01	0.03	0.86	0.15	< 0.03	<0.01	<0.02	3.0	1.1	0.35
			LF	3-0	4.6	<0.01	3.00	30.90	6.45	0.15	0.06	1.89	76.1	40.4	0.53
ED06	E.DYB	Nestow-AA	Ae	0-10	4.9	0.02	0.22	2.99	0.39	< 0.03	<0.01	0.11	6.6	3.6	0.55
			Bm1	10-25	4.8	<0.01	0.09	1.34	0.23	< 0.03	<0.01	<0.02	3.1	1.7	0.53
			LF	2-0	4.5	0.01	3.29	32.34	6.58	0.13	0.08	3.44	78.8	42.2	0.54
ED07	E.DYB	Nestow-AA	Ae	0-11	4.4	0.02	0.22	3.94	0.50	<0.03	<0.01	0.20	9.5	4.7	0.49
			Bm1	11-25	4.7	<0.01	0.09	1.37	0.32	<0.03	<0.01	<0.02	3.3	1.8	0.55
		Nestow-AA	LF	2-0	4.7	<0.01	0.69	15.67	1.82	0.05	0.03	0.61	35.8	18.2	0.51
ED08	E.DYB		Ahe/Ae	0-12	4.8	0.02	0.08	2.82	0.29	<0.03	<0.01	0.07	5.7	3.2	0.56
			Bm1	12-25	5.4	<0.01	0.04	1.24	0.21	<0.03	<0.01	<0.02	2.1	1.5	0.70
			LF	1-0	4.9	<0.01	1.12	24.75	4.36	<0.03	<0.01	0.70	49.5	30.2	0.61
ED09	E.DYB	Nestow-AA	Ae	0-6	4.7	0.01	0.13	3.09	0.51	<0.03	<0.01	0.08	6.9	3.7	0.54
ED06 E.DYB ED07 E.DYB ED08 E.DYB ED09 E.DYB ED10 E.DYB ED11 D.GL		Bm1	6-25	4.5	0.01	0.05	1.20	0.25	<0.03	<0.01	<0.02	3.8	1.5	0.40	
			LF	2-0	5.0	<0.01	2.56	38.83	8.81	0.16	0.03	3.11	75.4	50.2	0.67
ED10	E.DYB	Nestow-AA	Ae	0-7	5.6	0.01	0.26	7.05	0.97	<0.03	<0.01	0.11	10.5	8.3	0.79
			Bm1	7-25	4.6	0.01	0.14	1.48	0.27	<0.03	<0.01	0.02	5.4	1.9	0.36
FD11	D GI	Uncas	Ар	0-18	6.6	0.02	0.48	11.09	1.48	<0.03	<0.01	0.03	13.0	13.1	1.01
LDII	D.OE	Oncas	Bt1	18-25	6.7	0.04	0.26	9.52	2.07	<0.03	<0.01	< 0.02	12.9	11.9	0.92
			LF	13-0	6.5	0.06	2.54	87.91	10.44	<0.03	0.04	0.90	99.7	101.0	1.01
ED12	D.GL	Uncas		0-16	5.2	0.02	0.43	9.45	1.18	<0.03	<0.01	0.07	15.0	11.1	0.74
			AB	16-25	5.1	0.02	0.37	6.21	1.07	<0.03	<0.01	0.04	9.8	7.7	0.79
		A 1.	LFH	12-0	6.4	0.03	2.75	84.35	11.53	<0.03	<0.01	0.95	98.7	98.7	1.00
ED13	O.GL	Cooking Lake	Ae	0-13	5.4	0.01	0.43	5.26	1.15	<0.03	<0.01	0.06	9.0	6.9	0.76
		Lano	Bt1	13-25	5.6	0.02	0.60	10.90	3.06	<0.03	<0.01	<0.02	18.5	14.6	0.79
ED15	O.B	Peace Hills	Ар	0-25	5.7	<0.01	0.65	16.53	1.84	<0.03	<0.01	0.05	22.7	19.0	0.84

 Table B1. pH and Exchangeable Cation Data for Soils Sampled in the Edmonton East Grid Cell

² CEC – cation exchange capacity determined by buffered pH 7.0 ammonium acetate. BC – exchangeable base cations. Base Saturation – BEC/CEC

Site	Classification	Soil Series	Horizon	Depth	PH		Excha	angeable C	ations ar (cm	nd Cation nol kg⁻¹)	Exchan	ge Cap	acity		Base
				(cm)	(CaCl ₂)	Na	K	Ca	Mg	ĂÍ	Fe	Mn	CEC	BEC	Saturation
ED16	O.GL	Cooking	LFH	9-0	5.8	0.02	2.48	57.84	11.40	<0.03	<0.01	0.70	82.2	71.7	0.87
	O.GL	Lake	Bt	0-25	4.4	0.02	0.33	4.26	1.67	<0.03	<0.01	0.02	10.7	6.3	0.59
		Cooking	LF	9-0	5.5	0.14	2.42	53.21	9.44	< 0.03	<0.01	1.09	101.3	65.2	0.64
ED17	O.GL	Cooking Lake	Ahe/Ae1	0-12	4.1	0.27	0.41	5.08	1.66	0.04	<0.01	0.23	27.3	7.4	0.27
			Ae2	12-25	4.1	0.26	0.19	4.20	1.95	< 0.03	<0.01	0.04	14.2	6.6	0.47
		Cooking	LF	8-0	6.0	<0.01	1.30	35.43	4.48	< 0.03	<0.01	0.36	45.0	41.2	0.92
ED18	ED18 O.GL Cooking Lake	Ahe/Ae	0-21	5.3	0.02	0.38	7.89	0.94	<0.03	<0.01	0.09	12.7	9.2	0.73	
		Lako	Bt	21-25	4.8	0.04	0.52	11.69	1.78	< 0.03	<0.01	0.03	20.5	14.0	0.68
			LF	6-0	6.2	0.02	1.72	44.96	6.35	< 0.03	<0.01	0.57	54.8	53.1	0.97
ED19	O.DG	Helliwell-GR	Ahe/Ae	0-20	5.2	0.01	0.20	5.97	0.77	< 0.03	<0.01	0.07	10.6	6.9	0.66
			Bm1	20-25	5.1	0.01	0.13	2.30	0.39	< 0.03	<0.01	0.02	5.4	2.8	0.53
ED20	O.B	Ukalta	LFH	3-0	6.0	0.04	1.66	45.22	8.43	<0.03	0.01	0.08	62.1	55.4	0.89
LDZU	D20 O.B Okalia	Ukalla	Ahe	0-23	5.5	0.06	0.68	12.63	2.67	<0.03	<0.01	0.03	20.1	16.0	0.80
ED21	O.DG	Redwater	LF	5-0	5.9	0.04	1.79	67.24	9.30	<0.03	0.04	0.51	86.2	78.4	0.91
EDZI	0.00	Reuwalei	Ahe	0-25	5.8	0.02	0.48	7.07	1.21	< 0.03	<0.01	0.04	10.0	8.8	0.88
ED22	O.DG	Helliwell-GR	Ар	0-9	5.5	0.01	0.09	10.16	0.74	<0.03	<0.01	0.02	13.7	11.0	0.80
LDZZ	0.00		Btgj/Bm1	9-25	6.0	0.03	0.13	6.05	0.88	< 0.03	<0.01	<0.02	8.5	7.09	0.84
			LF	4-0	4.1	0.02	2.50	20.14	4.00	0.24	0.06	3.20	83.3	26.7	0.32
ED23	E.DYB	Nestow-AA	Ae	0-8	4.2	0.01	0.21	4.29	0.52	0.07	0.06	0.48	15.2	5.0	0.33
			Bm1	8-25	4.5	0.01	0.02	0.62	0.10	<0.03	<0.01	<0.02	2.0	0.8	0.37
ED24	O.DG	Helliwell	LFH	3-0	6.3	<0.01	5.80	56.49	10.27	<0.03	0.04	0.74	76.2	72.6	0.95
ED24	0.00	TIEIIIWEII	Ae	0-25	4.8	0.02	0.15	3.06	0.40	< 0.03	<0.01	0.03	6.7	3.6	0.54
ED25	O.DG	Helliwell	Ap/Ahe	0-25	5.4	<0.01	0.21	2.47	0.37	<0.03	<0.01	<0.02	4.7	3.0	0.65
			LF	5-0	4.1	0.02	0.73	14.63	2.27	0.18	0.10	1.30	40.6	17.7	0.43
ED26	E.DYB	Nestow-AA	Ae	0-4	4.5	0.01	0.07	3.81	0.30	<0.03	<0.01	0.16	9.2	4.2	0.45
			Bm1	4-32	5.2	0.02	0.05	1.51	0.22	< 0.03	<0.01	<0.02	3.7	1.8	0.48
TH1	O.B	Mundare	Ah	0-25	6.2	<0.01	0.30	4.40	0.70	nd	nd	nd	9.2	5.4	0.59
Brud	E.DYB	Nestow-AA	LFH	2-0	4.0	0.10	2.01	15.40	3.34	0.19	0.04	2.23	58.8	20.8	0.35
Biuu	E.UID	INCSIOW-AA	Ahe/Ae/Bm	0-25	4.7	0.01	0.09	1.64	0.29	0.02	0.01	0.15	4.2	2.0	0.49

Appendix Table B1. pH and Exchangeable Cation Data for Soils Sampled in the Edmonton East Grid Cell (concluded)

APPENDIX C: AREAS OF LAND SYSTEMS AND LAND COVER TYPES

Table C1. Areas of Land Systems and Land Cover Types

Land System	Total Area Cultivated			Grassla	nd	Trees	5	Shrub	s	Wetla	nd	Other L	and
Lanu System	(ha)	(ha)	%	(ha)	%	(ha)	%	(ha)	%	(ha)	%	(ha)	%
Bawlf Plain	5,197	3,287	63.25	1,765	33.96	142	2.73	0				3	0.05
Beaverhill Lake	20,835	965	4.63	6,259	30.04	1,369	6.57	0		12,159	58.36	83	0.40
Bruce Plain	37,029	12,592	34.00	22,530	60.84	1,748	4.72	0		17	0.05	143	0.39
Chipman Plain	44,270	23,420	52.90	17,236	38.93	3,410	7.70	0		69	0.16	134	0.30
Daysland Plain	1,360	852	62.66	449	32.98	54	3.95	0		0		5	0.40
Delph Upland	5,294	4,162	78.61	861	16.26	234	4.42	0		8	0.14	30	0.57
Edward Upland	1,675	14	0.84	592	35.34	1,031	61.55	0		0		38	2.27
Eldorena Plain	6,129	1,496	24.41	876	14.29	3,392	55.34	195	3.19	167	2.73	3	0.04
Ferlow Plain	9,601	4,526	47.14	2,469	25.72	2,428	25.29	0		152	1.58	27	0.28
Hairy Hill Plain	52	42	80.79	0		0		0		0		10	19.18
Hilliard Plain	95,250	71,526	75.09	15,702	16.48	7,639	8.02	146	0.15	221	0.23	17	0.02
Inland Plain	77,711	53,254	68.53	19,905	25.61	3,791	4.88	0		89	0.11	672	0.86
Irys Plain	1,268	535	42.21	637	50.26	46	3.59	0		0		50	3.94
Islet Upland	98,966	11,431	11.55	17,648	17.83	62,780	63.44	0		7,069	7.14	38	0.04
Kahwin Plain	30,075	25,464	84.67	3,569	11.87	759	2.52	3	0.01	142	0.47	139	0.46
Katchemut Upland	17,934	13,286	74.08	2,664	14.86	1,896	10.57	0		78	0.43	10	0.06
Little Beaver Plain	16,093	7,306	45.40	7,655	47.57	992	6.16	0		112	0.70	28	0.17
Lonestar Plain	620	229	36.87	329	53.09	28	4.57	0		0		34	5.48
Musidora Upland	19,323	4,540	23.50	7,269	37.62	7,400	38.29	22	0.12	78	0.40	14	0.07
Norma Plain	9,453	6,897	72.96	2,360	24.97	191	2.02	0		5	0.05	0	
North Saskatchewan River Valley	3,751	581	15.49	581	15.49	1,541	41.08	4	0.11	1,015	27.06	29	0.77
Pakan Plain	5,802	4,171	71.89	1,023	17.63	589	10.16	6	0.11	8	0.14	4	0.06
Partridge Plain	9,357	7,380	78.87	1,083	11.57	754	8.06	0		39	0.42	101	1.08
Redclay Plain	2,567	1,798	70.04	670	26.09	68	2.66	2	0.08	20	0.79	9	0.34
Redwater Plain	8,068	3,577	44.33	1,942	24.07	2,354	29.17	124	1.54	71	0.88	0	
Royal Park Plain	14,636	11,530	78.78	2,544	17.38	538	3.67	0		7	0.05	18	0.12
Ryley Plain	161,012	77,711	48.26	69,489	43.16	11,904	7.39	0		666	0.41	1,242	0.77
Thorhild Plain	91	55	60.71	8	8.56	28	30.73	0		0		0	0.00
Vegreville Plain	5,837	2,735	46.86	2,671	45.77	266	4.55	0		0		164	2.81
Watt Lake Plain	21,190	14,188	66.96	5,883	27.76	424	2.00	1	0.01	632	2.98	62	0.29
Whitford Plain	8,102	6,350	78.37	1,249	15.42	421	5.20	0		0		83	1.02
Not Accounted for	44	18	41.96	12	26.19	12	27.00	1	1.45	1	3.23	0	0.17
Total	738,596	375,918		217,930		118,229		504		22,825		3,190	

APPENDIX D: LAND AREA ESTIMATION OF SENSITIVITY CLASSES

LAND AREA ESTIMATION OF SENSITIVITY CLASSES

Five Land Systems in the Edmonton East grid cell were found to include Soil Series to be Sensitive or Moderately Sensitive to acidification according to the ARC model. The Primula (Eluviated Eutric Brunisol) and Nestow (Eluviated Dystric Brunisol) are considered together as being Sensitive to Moderately Sensitive to acidification. The Helliwell (Orthic Dark Gray Chernozem) and Mundare (Orthic Black Chernozem) Soil Series are rated as being of Moderate to Low sensitivity. The table below indicates the sensitivity classes of these soils, as well their estimated proportions in five Land Systems. The assignment of proportions was as follows:

- Land System with a major soil and two minor soils: the major soil is estimated to constitute 70% of the Land System, and the minor soils are estimated to constitute 15% each.
- Land System with two major soil series and two minor soil series: the major soils are estimated to each constitute 35% of the Land System, and the minor soils are estimated to constitute 15% each.

Soil Series	Mundare	Primula (Nestow)	Peace Hills	Redwater	Helliwell	Uncas	Angus Ridge	Camrose	Manatokan, Misc. Organics
Series Sensitivity	M-L	S-M	L	L	M-L	L	L	L	L
Edward Upland		35%		35%		15%			15%
Eldorena Plain	15%	35%	35%						15%
Musidora Upland		15%	35%		35%				15%
Redwater Plain	70%	15%	15%						
Norma Plain	15%		15%				35%	35%	

The above percentages of Soil Series in the Land Systems were then compared to the land cover data (Appendix C). It was assumed that the cultivated lands were occupied by the soils with the highest agricultural capability, and that land with shrub, tree and grassland cover would have the sandiest soils, namely Mundare, Helliwell and Primula/Nestow. Also, percent areas of each of the occurrences of Primula//Nestow, Helliwell and Mundare were halved and assigned to two sensitivity classes because of their dual ratings. For example, Primula/Nestow is rated Sensitive to Moderately Sensitive; therefore, half their areas were assigned to each of these sensitivity ratings. Details of the rating derivations are provided below for the five Land Systems with Moderate and Sensitive ratings.

Musidora Upland

- Cultivated 24%; not rated for sensitivity; assume soil is Peace Hills 24%;
- Other Lands <1%; not rated for sensitivity
- Grassland , Shrubs and Trees 76%; assume soils are 35% Helliwell; 15% Primula; remaining 11% of Peace Hills; 15% Mantokan
- Not rated <1%
- Sensitive Half of Primula (7-8%)

- Moderate Sensitivity Half of Primula (7-8%) and half of Helliwell (17-18%);
- Low Sensitivity half of Helliwell (17-18%); Peace Hills (11%), Manatokan 15%
- Summary: Sensitive 7-8%; Moderate 24-25%; Low 44%; Not Rated 24%

Norma Plain

- Cultivated 73%; assume Angus Ridge, Peace Hills and Camrose soils
- Grassland, Shrubs and Trees 27%; assume all of Mundare soils (15%) and 12% of Camrose soils (the most severe Solonetzic soils are assumed to be under grass; reconnaissance of the area during soil sampling suggested that this was indeed the case).
- Moderate Sensitivity Half of Mundare (7-8%);
- Low Sensitivity Half of Mundare (7-8%), and all other soils
- Summary: Moderate 7-8%; Low 17%; Not Rated 75%

Edward Upland

- Cultivated 1%; Redwater assumed 1%
- Other 2%; not rated
- Grassland , Shrubs and Trees 97%; 34% Redwater; 35% Primula; 15% Uncas; 15% Mantokan/Organics
- Not rated 1%
- Sensitive Half of Primula (17-18%)
- Moderate Sensitivity Half of Primula (17-18%)
- Low Sensitivity Mundare (34%), Uncas 15%, Manatokan 15%
- Summary: Sensitive 17-18%; Moderate 17-18%; Low 64%

Redwater Plain

- Cultivated 44%; Peace Hills 15%; Mundare 29%
- Other 2%; not rated
- Grassland, Shrubs and Trees 54%; 15% Primula; 39% Mundare
- Sensitive Half of Primula (7-8%)
- Moderate Sensitivity Half of Primula (7-8%) and half of Mundare (19-20%)
- Low Sensitivity Half of Mundare (20%)
- Summary: Sensitive 7-8%; Moderate 28%; Low 19-20%; Not Rated 46%

Eldorena Plain

- Cultivated 24%; Peace Hills 24%
- Other <1%; not rated
- Grassland , Shrubs and Trees 76%; 11% Peace Hills; 35% Primula; 15% Mundare; 15% Mantokan
- Sensitive Half of Primula (17-18%)
- Moderate Sensitivity Half of Primula (17-18%) and half of Mundare (7-8%);
- Low Sensitivity Half of Mundare (7-8%), Peace Hills, non-cultivated 11%, Manatokan 15%
- Summary: Sensitive 17-18%; Moderate 24%; Low 34%; Not Rated 25%

BASELINE SOIL AND TERRAIN MAPS